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2 **UNIVERSIDADE FEDERAL RURAL DE PERNAMBUCO**
3 **PRÓ-REITORIA DE PESQUISA E PÓS-GRADUAÇÃO**
4 **PROGRAMA DE PÓS-GRADUAÇÃO EM RECURSOS PESQUEIROS E AQUICULTURA**
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7 **UTILIZAÇÃO DE MODELOS PARA DADOS LIMITADOS NA AVALIAÇÃO**
8 **DOS ESTOQUES DAS PRINCIPAIS ESPÉCIES DE LUTJANIDAE**
9 **CAPTURADAS NO NORDESTE**
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14 **Andrey Paulo Cavalcanti Soares**
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20 Pós-Graduação em Recursos
21 Pesqueiros e Aquicultura da
22 Universidade Federal Rural de
23 Pernambuco como exigência para
24 obtenção do título de Doutor.
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À minha família, que foi meu baluarte enquanto eu aprendia que a ciência, assim
como a vida, é feita de tentativa, erro e persistência.

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Resumo

214
215 As pescarias artesanais do Nordeste do Brasil caracterizam-se, em sua maioria,
216 por limitações na disponibilidade e na qualidade de dados, o que dificulta a aplicação de
217 métodos tradicionais de avaliação de estoques pesqueiros. Nesse contexto, os modelos
218 para dados limitados (MDL) têm se mostrado alternativas promissoras para subsidiar o
219 manejo, permitindo estimar o estado dos estoques mesmo sob cenários de escassez de
220 informações. Esta tese teve como objetivo geral avaliar a utilidade de modelos para dados
221 limitados, em especial o Stock Assessment Continuum (SAC) Tool, na estimativa do
222 estado dos estoques das principais espécies de Lutjanidae exploradas no Nordeste do
223 Brasil, além de gerar subsídios técnicos para o manejo pesqueiro. No Capítulo I, foi
224 realizada uma avaliação metodológica do desempenho do SAC Tool por meio da
225 reanálise do estoque de *Ocyurus chrysurus*, previamente avaliado no âmbito do programa
226 REVIZEE com métodos tradicionais baseados em idade. Diferentes cenários foram
227 testados, combinando dados de captura, composições de comprimento e informações de
228 história de vida, incluindo configurações integradas e abordagens baseadas
229 exclusivamente em comprimento. Os resultados mostraram que, apesar das diferenças na
230 quantidade e no tipo de dados utilizados, todos os cenários convergiram para a mesma
231 conclusão qualitativa: o estoque encontra-se sobreexplorado e submetido à sobrepesca.
232 Os modelos integrados apresentaram maior precisão, enquanto os cenários baseados
233 apenas em comprimento exibiram maior incerteza, como esperado, mas mantiveram a
234 robustez do sinal de depleção do estoque. Esses resultados indicam que os MDL, quando
235 adequadamente especificados, são capazes de reproduzir padrões consistentes com
236 avaliações tradicionais, mesmo sob forte redução de dados. No Capítulo II, os modelos
237 para dados limitados foram aplicados para atualizar a avaliação dos estoques de quatro
238 espécies-chave de lutjanídeos no Nordeste do Brasil: *Lutjanus analis*, *Lutjanus jocu*,
239 *Lutjanus synagris* e *Ocyurus chrysurus*. Foram utilizados dados históricos de captura,
240 composições de comprimento, índices de abundância e parâmetros de história de vida
241 para estimar o estado dos estoques em termos de biomassa relativa e mortalidade por
242 pesca em relação aos pontos de referência biológicos. De modo geral, os resultados
243 indicaram que a maioria dos estoques apresenta sinais de depleção e níveis de exploração
244 acima do sustentável, reforçando a necessidade de medidas de manejo mais
245 conservadoras e de ações voltadas à recuperação dos estoques. Em conjunto, os resultados
246 desta tese demonstram que os MDL constituem ferramentas robustas e operacionalmente
247 viáveis para a avaliação de estoques em contextos de escassez de informações, como os

248 observados nas pescarias artesanais do Nordeste do Brasil. Além de validarem a utilidade
249 do SAC Tool, os resultados fornecem evidências de que essas abordagens podem gerar
250 diagnósticos confiáveis sobre o estado dos estoques e apoiar a implementação de
251 estratégias de manejo precaucionárias. Assim, os MDL se apresentam como componentes
252 centrais de um sistema de gestão adaptativo, no qual a incorporação gradual de novos
253 dados tende a aprimorar a precisão das avaliações e a sustentabilidade das pescarias.

Abstract

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Small-scale fisheries in Northeast Brazil are largely characterized by limitations in data availability and quality, which restrict the application of traditional stock assessment methods. In this context, data-limited models (DLMs) have emerged as promising alternatives to support fisheries management, allowing the estimation of stock status even under severe data scarcity. The main objective of this thesis was to evaluate the usefulness of data-limited models, particularly the Stock Assessment Continuum (SAC) Tool, for estimating the status of the main Lutjanidae stocks exploited in Northeast Brazil, and to provide technical support for fisheries management. In Chapter I, a methodological evaluation of the performance of the SAC Tool was carried out through the reanalysis of the *Ocyurus chrysurus* stock, previously assessed under the REVIZEE program using traditional age-based approaches. Different scenarios were tested by combining catch data, length compositions, and life-history information, including integrated configurations and length-only approaches. The results showed that, despite differences in the quantity and type of data used, all scenarios converged to the same qualitative conclusion: the stock is overfished and subject to overfishing. Integrated models produced more precise estimates, whereas length-only scenarios displayed higher uncertainty, as expected, but consistently retained the robust signal of stock depletion. These findings indicate that, when properly specified, data-limited models are able to reproduce conclusions consistent with traditional assessments, even under substantial data reduction. In Chapter II, DLMs were applied to update the stock assessments of four key lutjanid species in Northeast Brazil: *Lutjanus analis*, *Lutjanus jocu*, *Lutjanus synagris*, and *Ocyurus chrysurus*. Historical catch data, length compositions, abundance indices, and life-history parameters were used to estimate stock status in terms of relative biomass and fishing mortality with respect to biological reference points. Overall, the results indicate that most stocks show clear signs of depletion and fishing pressure above sustainable levels, reinforcing the need for more conservative management measures and actions aimed at stock rebuilding. Taken together, the results of this thesis demonstrate that data-limited models represent robust and operationally feasible tools for stock assessment in data-poor contexts such as the small-scale fisheries of Northeast Brazil. In addition to validating the applicability of the SAC Tool, this study provides evidence that these methods can generate reliable diagnoses of stock status and effectively support precautionary management strategies. Therefore, data-limited models should be viewed as central components of an adaptive management framework, in which the progressive

288 incorporation of new data is expected to improve assessment accuracy and contribute to
289 the long-term sustainability of fisheries.
290

291 **Introdução geral**

292 A humanidade vem consumindo recursos naturais de forma consistente ao longo
293 da história. Entretanto, a Revolução Industrial representou um ponto de inflexão nesse
294 processo: a população mundial saltou de pouco mais de 1 bilhão em 1800 para 6 bilhões
295 em 2000 e ultrapassou 8 bilhões em 2021 (UN-DESA, 2022), intensificando de maneira
296 sem precedentes a demanda por alimentos e matérias-primas. Como resultado direto desse
297 crescimento, aliado aos avanços tecnológicos, observou-se a ampliação de alterações
298 ambientais em escala global, incluindo a destruição de habitats e mudanças climáticas,
299 frequentemente sem a consideração de limites de sustentabilidade.

300 Até o século XIX, a pesca era uma atividade relativamente pequena, restrita à
301 extração limitada de organismos e associada a baixo impacto ambiental. As evoluções
302 tecnológicas observadas no período pós-Segunda Guerra Mundial ampliaram
303 significativamente o poder de captura em escala global e, particularmente, no Brasil
304 (Paes, 2002). A partir da década de 1970, a produção pesqueira marinha passou a
305 apresentar sinais de estabilização, enquanto os desembarques de espécies-chave, como
306 sardinha e camarão, entraram em declínio (Pauly *et al.*, 2002). A extração acima da
307 capacidade de reposição altera a estrutura populacional, reduz recrutamento e capacidade
308 de regeneração dos estoques, podendo levar a colapsos locais (FAO, 1996; Casey e
309 Myers, 1998; Dulvy *et al.*, 2004).

310 Espécies com crescimento lento, maturação tardia e distribuição restrita, como
311 diversos representantes da família Lutjanidae, são particularmente vulneráveis à
312 sobrepesca. A intensificação da exploração tende a: (i) reduzir o contingente de
313 indivíduos adultos, refletida na diminuição do comprimento médio das capturas; e (ii)
314 antecipar a captura para tamanhos menores, comprometendo o recrutamento e a reposição
315 populacional (Sadovy, 2001; Paes, 2002; Dulvy *et al.*, 2004). Em escala global, a
316 proporção de estoques biologicamente sustentáveis caiu de 90% para 65,8% entre 1974 e
317 2017 (FAO, 2020).

318 Apesar da crescente pressão sobre os recursos pesqueiros, a maioria das pescarias
319 é classificada como limitadas em dados, tanto em termos de qualidade quanto de
320 disponibilidade de informações, o que dificulta avaliações de estoques robustas. Diante
321 da escassez de dados típica das pescarias artesanais do Nordeste, modelos para dados
322 limitados (MDL) tornam-se ferramentas essenciais para fornecer diagnósticos
323 minimamente robustos sobre o estado dos estoques. As seções seguintes descrevem: (1)

324 a evolução da pesca no Brasil, (2) o papel dos lutjanídeos e (3) a relevância dos modelos
325 para dados limitados.

326 *Exploração pesqueira no Brasil*

327 Ao longo do último século, a atividade pesqueira no Brasil passou de uma
328 prática predominantemente artesanal e de subsistência para um setor
329 progressivamente industrializado e diversificado, assumindo papel relevante na
330 economia nacional. Esse processo de expansão esteve associado a desafios
331 ecológicos, econômicos e sociais, refletindo mudanças tecnológicas, transformações
332 no manejo dos recursos e novas demandas socioeconômicas.

333 *Da pesca artesanal à industrialização (1900-1970)*

334 Até meados do século XX, a pesca no Brasil era essencialmente artesanal,
335 voltada ao consumo local (Prá e D'Agostini, 2023). A organização institucional do
336 setor teve início com a promulgação do Estatuto da Confederação Geral dos
337 Pescadores, em 1950. Ainda na década, o processo de industrialização do país
338 impulsionou a modernização da frota pesqueira com a introdução de embarcações
339 motorizadas, tecnologias de detecção sistemas de conservação, ampliando
340 significativamente a capacidade de captura (Abdallah e Bacha, 1999; Santos *et al.*,
341 2012).

342 A criação da Superintendencia do Desenvolvimento da Pesca, em 1962, e o
343 Decreto-Lei 221/1967, que instituiu subsídios fiscais e isenções de combustível,
344 favoreceram a expansão da frota industrial e da capacidade produtiva. Esse período
345 foi marcado por aumento expressivo do esforço pesqueiro e da exploração dos
346 recursos marinhos, sem o desenvolvimento proporcional de mecanismos de
347 monitoramento e controle (Neiva, 1990).

348 *Estabilização e conflitos (1970-1990)*

349 Na década de 1970, tornaram-se evidentes sinais de sobre-exploração com a
350 estagnação da produção marinha e declínio de estoques tradicionais como a
351 sardinha-verdadeira e o camarão. A ausência de delimitação clara de áreas de pesca e
352 o monitoramento deficiente favoreceram práticas predatórias (Paez, 1993; Abdallah
353 e Bacha, 1999). Nesse contexto, os conflitos entre a pesca industrial e artesanal se
354 intensificaram, uma vez que frotas mais capitalizadas passaram a competir

355 diretamente com comunidades costeiras dependentes dos recursos pesqueiros
356 (Diegues, 1995).

357 *Transição para a sustentabilidade (1990-2007)*

358 Na década de 1990, a crescente percepção sobre os limites da exploração
359 pesqueira levou à adoção de políticas voltadas para a sustentabilidade. O Brasil
360 ratificou acordos internacionais, como o Código de Conduta para a Pesca Responsável
361 da FAO (FAO, 1995), e passou a implementar instrumentos de gestão, incluindo
362 períodos de defeso, cotas de captura e áreas protegidas. Entretanto, a efetividade
363 dessas medidas foi limitada pela falta de dados confiáveis e contínuos. Programas de
364 monitoramento, como o ESTATPESCA (IBAMA, 2008a), tiveram papel relevante,
365 mas sua interrupção em 2007 resultou em uma lacuna crítica na gestão dos recursos.

366 *Atualidade (2008-2024)*

367 Nas últimas duas décadas, o setor pesqueiro brasileiro tem enfrentado desafios
368 persistentes relacionados à sobrepesca, à degradação ambiental e às mudanças
369 climáticas. As pescarias artesanais seguem como componente central da economia
370 pesqueira, especialmente no Nordeste e na Amazônia, mas operam sob crescente
371 pressão sobre os estoques (Viana, 2021a, 2021b). Nesse período, a ausência de
372 monitoramento contínuo e de dados sistemáticos limitou a efetividade da gestão. Com
373 o reestabelecimento do Ministério da Pesca e Aquicultura (MPA), houve a retomada
374 da estatística pesqueira nacional, o que resultou na publicação do Boletim Estatístico
375 da Pesca e Aquicultura 2023–2024, e na adoção de políticas públicas voltadas ao
376 setor, como o Plano Safra da Pesca e Aquicultura (MPA, 2022) e o Plano Nacional da
377 Pesca Artesanal (MPA, 2024).

378 *Lutjanídeos*

379 Os lutjanídeos, comumente chamados de “vermelhos”, compõem parcela
380 relevante das capturas demersais tropicais e possuem alto valor de mercado. Os
381 lutjanídeos são espécies marinhas tropicais e subtropicais encontradas em diversas
382 partes do mundo e possuem um alto valor socioeconômico comercial e recreativo
383 (Grimes 1987, Duarte-Garcia 1999, Freitas et al 2011, Resende et al. 2003). São
384 espécies que desempenham papel ecológico fundamental nos ecossistemas recifais,
385 atuando como predadores de topo que regulam as populações de presas, incluindo
386 peixes menores e invertebrados (Fernandes et al. 2020; Freitas et al. 2011; Monteiro

387 et al. 2009; Moreno-Sanchez et al. 2016; Valle-Lopez et al. 2021). São agrupados com
388 espécies ecologicamente semelhantes, como os meros, compondo o “snapper-grouper
389 complex”, caracterizados por baixas taxas de crescimento, maturação tardia e
390 formação de agregações reprodutivas, fatores que aumentam sua vulnerabilidade à
391 pesca intensiva (Chuenpagdee e Pauly, 2005; Schärer-Umpierre *et al.*, 2014; França
392 *et al.*, 2021). Além de seu papel direto como controladores de biomassa, essas
393 espécies participam de fluxos tróficos que conectam habitats costeiros e recifais,
394 transportando energia e nutrientes entre diferentes compartimentos do ecossistema
395 (Derviche et al., 2025). Alterações nas populações dessas espécies, seja por
396 sobrepesca ou perda de habitat, podem gerar efeitos negativos em cascata,
397 comprometendo a resiliência e a capacidade de recuperação dos recifes. Por isso, o
398 manejo sustentável desses predadores é não apenas uma questão de conservação das
399 espécies-alvo, mas também de manutenção das funções ecológicas essenciais dos
400 recifes tropicais.

401 No Brasil, particularmente na Região Nordeste, os lutjanídeos são
402 historicamente explorados e representam importante fonte de renda para comunidades
403 de pesca artesanal, tanto pelo volume de desembarques quanto pelo elevado valor
404 comercial (Resende, Ferreira e Frédou, 2003). Ao todo, doze espécies de lutjanídeos
405 são alvo de pescarias na região, com desembarques que aumentaram de
406 aproximadamente 4.000 toneladas em 1990 para mais de 10.500 toneladas no ano de
407 2006, gerando receita bruta superior a 58 milhões de reais nesse ano (IBAMA, 2008b;
408 Freire et al, 2021). No início dos anos 2000, o programa REVIZEE avaliou os
409 principais estoques explorados na Zona Econômica Exclusiva (ZEE), incluindo os
410 lutjanídeos do Nordeste, utilizando métodos tradicionais de avaliação baseados em
411 idade, como análise de população virtual e modelos de rendimento por recruta. Essas
412 avaliações indicaram, de forma consistente, estoques sobre-explorados ou próximos
413 do limite máximo de exploração, sem que medidas efetivas de manejo fossem
414 implementadas (Frédou, Ferreira e Letourneur, 2009a, 2009b; Lessa, Bezerra-Júnior
415 e Nóbrega, 2009). A interrupção do programa nacional de estatística pesqueira em
416 2007 fragmentou as séries históricas de desembarque, especialmente para espécies
417 recifais capturadas pela frota artesanal, agravando a limitação de dados disponível.
418 Entre as espécies avaliadas pelo REVIZEE, quatro respondem pela maior parte dos
419 desembarques regionais e constituem o foco deste estudo: *Lutjanus analis*, *L. jocu*, *L.*
420 *synagris* e *Ocyurus chrysurus*.

421 *Lutjanus analis*

422 A cioba (*Lutjanus analis*) distribui-se pelo Atlântico Ocidental, do Caribe ao
423 sudeste do Brasil, ocupando ambientes costeiros associados principalmente a recifes
424 e fundos consolidados (Allen, 1985). Os indivíduos atingem comumente 40–50 cm,
425 podendo alcançar cerca de 1 m, com maturidade sexual próxima de 36 cm e
426 longevidade estimada em até 29 anos (Lessa, Bezerra-Júnior e Nóbrega, 2009). Os
427 desembarques foram modestos nas décadas iniciais da série histórica, mas
428 aumentaram progressivamente a partir dos anos 1990, atingindo o pico em 2005, com
429 cerca de 3,3 mil toneladas. Após esse período, observa-se redução nas capturas em
430 relação ao pico histórico, indicando mudança no padrão de desembarques da espécie.

431 *Lutjanus jocu*

432 O dentão (*Lutjanus jocu*) apresenta ampla distribuição no Atlântico Ocidental,
433 ocupando principalmente ambientes recifais e fundos rochosos costeiros, enquanto
434 juvenis ocorrem em áreas litorâneas e estuarinas. A espécie atinge frequentemente
435 35–45 cm, podendo ultrapassar 1 m, com maturidade próxima de 32 cm e longevidade
436 estimada em até 29 anos (Allen, 1985; Lessa, Bezerra-Júnior e Nóbrega, 2009). Os
437 desembarques aumentaram gradualmente a partir do final dos anos 1990, alcançando
438 máximo em 2003, com cerca de 2,1 mil toneladas. A partir de 2004, observa-se
439 redução nos desembarques em relação ao período de maior captura.

440 *Lutjanus synagris*

441 O ariocó (*Lutjanus synagris*) ocorre ao longo da costa tropical do Atlântico
442 Ocidental, associado a recifes e fundos arenosos vegetados. Os indivíduos medem
443 comumente entre 20 e 30 cm, podendo alcançar 65 cm, atingindo maturidade por volta
444 de 18 cm e vivendo até cerca de 22 anos (Lessa, Nóbrega e Bezerra-Júnior, 2004;
445 Freitas et al., 2011). Os desembarques apresentaram crescimento contínuo ao longo
446 da série histórica, atingindo pico em 2008, com aproximadamente 2,7 mil toneladas,
447 seguido por estabilização das capturas em níveis inferiores ao pico observado.

448 *Ocyurus chrysurus*

449 A guaiúba (*Ocyurus chrysurus*) distribui-se amplamente no Atlântico
450 Ocidental, sendo comum em ambientes recifais costeiros. Os indivíduos apresentam
451 comprimento médio entre 30 e 40 cm, podendo atingir até 90 cm, com maturidade
452 sexual próxima de 20 cm e longevidade estimada em cerca de 19 anos (Lessa,

453 Bezerra-Júnior e Nóbrega, 2009). Essa espécie apresentou os maiores volumes de
454 desembarque entre os lutjanídeos avaliados, alcançando pico em 2004, com cerca de
455 7,5 mil toneladas. A partir de 2005 observa-se redução contínua nos desembarques
456 em relação ao pico histórico registrado.

457 *Modelos para dados limitados*

458 De modo geral, o manejo pesqueiro busca conciliar a utilização dos recursos
459 em benefício da sociedade com a mitigação dos impactos ambientais associados à
460 atividade pesqueira (Sainsbury, Punt e Smith, 2000; Garcia *et al.*, 2003). Nesse
461 sentido, a avaliação de estoques busca fornecer aos tomadores de decisão informações
462 sobre o equilíbrio entre exploração e manutenção dos recursos ao longo do tempo
463 (Sparre e Venema, 1998).

464 De forma simplificada, uma avaliação de estoques busca responder a três
465 questões centrais: a escala do estoque, o estado em relação a pontos de referência e a
466 sua produtividade. A escala se refere à quantificação absoluta do estoque, como
467 biomassa total, biomassa desovante. O estado representa a condição atual do estoque
468 em relação a referências biológicas, por exemplo a razão entre a biomassa atual e a
469 biomassa no rendimento máximo sustentável (B/B_{MSY}) ou a biomassa virgem (B/B_0).
470 Já a produtividade envolve processos que determinam a capacidade de reposição
471 populacional, incluindo parâmetros como a taxa intrínseca de crescimento (r),
472 mortalidade natural (M) e *steepness* (h), que refletem a resiliência da espécie à
473 exploração (Cope, 2024b). Entretanto, avaliações robustas dependem da
474 disponibilidade de séries históricas confiáveis de abundância, idade ou comprimento,
475 informação ausente em grande parte das pescarias mundiais, especialmente em países
476 em desenvolvimento (Dowling *et al.*, 2019; Reis, 1992; Frédou, 2004). Diante desse
477 cenário, modelos para dados limitados (MDLs) vêm sendo desenvolvidos para
478 maximizar a informação disponível e permitir inferências mesmo sob elevada
479 incerteza (Chrysafi e Kuparinen, 2015; Dowling *et al.*, 2019; Cope *et al.*, 2023; Cope,
480 2024b).

481 Entre essas abordagens, destacam-se os modelos baseados em comprimento,
482 que utilizam composições de comprimento e parâmetros de história de vida para
483 inferir níveis de exploração e estado do estoque. Métodos baseados em comprimento
484 assumem que a mortalidade por pesca se reflete na estrutura de tamanhos e no
485 comprimento médio das capturas (Pons, Cope e Kell, 2020). Sob essa perspectiva,

486 esses modelos utilizam composições de comprimento para estimar a mortalidade e o
487 estado dos estoques pesqueiros. Como a coleta de dados de comprimento é
488 relativamente simples e de baixo custo, esse costuma ser o principal tipo de
489 informação disponível e, em situações de limitação de dados, pode representar a única
490 fonte de informação para avaliação (Hordyk et al., 2015a; Mildenberger, Taylor e
491 Wolff, 2017). Os principais pressupostos assumidos pelos modelos baseados em
492 comprimento incluem: (1) a representatividade das composições de comprimento
493 amostradas, (2) o estoque se encontra em condição próxima ao equilíbrio, (3) que os
494 parâmetros de história de vida são constantes e representativos da população avaliada
495 e (4) os padrões de seletividade e maturidade podem ser descritos por curvas
496 logísticas. Entre os métodos mais utilizados destacam-se o Length-based Bayesian
497 Biomass (LBB; Froese et al., 2018), que permite estimar o estado do estoque a partir
498 exclusivamente de composições de comprimento, e o Length-Based Spawning
499 Potential Ratio (LBSPR, Hordyk et al., 2015a), amplamente empregado para inferir o
500 potencial reprodutivo remanescente sob diferentes níveis de exploração.

501 Outra classe de modelos utiliza séries temporais de captura associadas a
502 parâmetros biológicos para estimar níveis sustentáveis de exploração. Métodos
503 baseados em captura permitem inferir produtividade, biomassa relativa e níveis
504 sustentáveis de captura mesmo na ausência de dados independentes de abundância.
505 Para esses modelos, os principais pressupostos são que as séries de captura sejam
506 representativas da biomassa e que o esforço pesqueiro permaneça relativamente
507 estável ao longo do tempo, de modo que variações nas capturas reflitam
508 principalmente mudanças na biomassa do estoque, desconsiderando possíveis
509 alterações na dinâmica e no comportamento da frota. Entre esses métodos, destaca-se
510 o CMSY++ (Froese et al., 2021), amplamente aplicado em pescarias com séries
511 históricas de captura limitadas (MPA, 2022; Chen, 2024; Cardoso et al, 2026).

512 Além dessas abordagens, modelos integrados combinam diferentes fontes de
513 informação, como capturas, composições de comprimento e parâmetros biológicos
514 dentro de uma estrutura analítica. Ao integrar diferentes fontes de dados e
515 metodologias, modelos integrados tendem a aumentar a confiabilidade das
516 estimativas, especialmente em contextos de alta variabilidade ou limitação de dados.
517 Essa flexibilidade permite construir, dentro de uma mesma estrutura de modelagem,
518 configurações que variam desde abordagens simplificadas até avaliações mais
519 completas, reduzindo a lacuna entre métodos tradicionalmente considerados data-rich

520 e contextos operacionais caracterizados por escassez de informação. Entre essas
521 abordagens destaca-se o Stock Assessment Continuum Tool (Cope, 2024b),
522 empregado neste estudo para integrar diferentes fontes de dados sob cenários de
523 informação limitada.

524 Embora a gestão pesqueira busque otimizar o uso de recursos e mitigar
525 impactos (Sainsbury, Punt e Smith, 2000), séries históricas completas são raras em
526 pescarias artesanais (Cope et al. 2023). Neste contexto, os MDLs são alternativas
527 pragmáticas aos métodos tradicionais. A maioria dos estoques pesqueiros no Nordeste
528 do Brasil, incluindo os lutjanídeos, possuem dados limitados e, portanto, necessitam
529 de métodos alternativos de avaliação. Além disso, existe a necessidade de atualização
530 do estado dos estoques dessas espécies, que são espécies de grande importância
531 ecológica e socioeconômica no Brasil. Dessa forma, esta tese busca avaliar a eficácia
532 dos MDLs em relação aos métodos tradicionais, validando sua utilidade em pescarias
533 demersais de pequena escala no Nordeste brasileiro e atualizando o estado dos
534 estoques das principais espécies de lutjanídeos da região, fornecendo subsídios para
535 o manejo sustentável e aplicação dessas abordagens a pescarias com características
536 semelhantes.

537 **Objetivos**

538 *Geral*

539 Avaliar o desempenho e a robustez de modelos para dados limitados na estimativa
540 do estado dos estoques das principais espécies de lutjanídeos exploradas no Nordeste do
541 Brasil, visando subsidiar o manejo pesqueiro em contextos de escassez de dados.

542 *Específicos*

- 543 • Avaliar o desempenho de modelos para dados limitados sob diferentes níveis
544 de disponibilidade de informação, comparando suas estimativas com
545 avaliações históricas baseadas em métodos tradicionais, a fim de investigar a
546 robustez dos diagnósticos de estado do estoque.
- 547 • Aplicar modelos adequados a dados limitados para estimar o estado atual dos
548 estoques das quatro principais espécies de lutjanídeos exploradas no Nordeste
549 do Brasil, analisando padrões de depleção, mortalidade por pesca e a incerteza
550 associada às estimativas.

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- Integrar os resultados metodológicos e aplicados para discutir implicações para o manejo de pescarias demersais tropicais, identificando potencialidades, limitações e diretrizes para adoção desses métodos em sistemas pesqueiros com dados escassos.

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Capítulo I

Artigo: “Information content and stock status: comparing traditional and contemporary assessment approaches”

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596 **INFORMATION CONTENT AND STOCK STATUS: COMPARING TRADITIONAL**
597 **AND CONTEMPORARY ASSESSMENT APPROACHES**

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605 *Abstract*

606 Data-limited methods are increasingly used in tropical fisheries where traditional
607 stock assessments are often unfeasible. Here, we evaluate the performance of the Stock
608 Assessment Continuum (SAC) Tool by reassessing the early-2000s evaluation of *Ocyurus*
609 *chrysurus* carried out under the REVIZEE program. Six scenarios incorporating different
610 combinations of catch and length compositions were compared to the original age-
611 structured Virtual Population Analysis, which served as a benchmark for validation.
612 Across all scenarios evaluated, the SAC Tool indicated that the stock was overfished and
613 experiencing overfishing, with this signal remaining robust despite differences in data
614 availability and model assumptions. Model performance varied with data availability:
615 integrated scenarios produced more precise estimates, whereas length-only approaches
616 exhibited wider uncertainty but retained the same qualitative signal of depletion. These
617 results show that key stock status signals remain robust across modern data-limited
618 approaches, despite major reductions in data inputs and alternative model assumptions,
619 supporting their application in small-scale demersal fisheries. Continued improvement of
620 data collection and refinement of model assumptions will further strengthen the reliability
621 of these approaches.

622 **Keywords:** Data-limited stock assessment, length-based methods, integrated
623 models

624 *Introduction*

625 As with all human activities, fishing has an impact on the environment, whether
626 through pollution, habitat disturbance, or intense fishing pressure, which can result in
627 stock depletion. Globally, several fish species of high socio-economic importance have
628 been classified as overfished, with the proportion of assessed fish stocks considered
629 biologically sustainable declining from 90% to 65.8% between 1974 and 2017, while the
630 proportion exceeding sustainability thresholds increased from 10% to 34.2% over the
631 same period (FAO, 2020). However, these global estimates are derived primarily from
632 stocks with formal assessments, whereas a substantial proportion of fisheries worldwide
633 lack sufficient data to support routine stock status evaluation. Indeed, most fisheries are
634 classified as data-limited, indicating that there is a major issue concerning data
635 availability or quality (FAO et al., 2023). This gap highlights the need for assessment
636 approaches suitable for data-limited fisheries.

637 In the past, assessments of data-limited fish stocks were constrained by the
638 limitations of the data available and were unable to provide reference points such as
639 maximum sustainable yield, total allowable catch, or maximum effort at the sustainable
640 yield (Geromont and Butterworth, 2015; Sagarese et al., 2019). Recently, data-limited
641 models were developed as a subset of stock assessment modelling approaches that attempt
642 to maximize the use of the available data despite its limitations (Chrysafi and Kuparinen,

2015; Cope et al., 2023; Cope, 2024a; Dowling et al., 2019). For instance, when abundance indices and age-composition data are unavailable but length compositions and catch data are available, data-limited methods may provide a more suitable option (Rudd et al. 2021). These models usually have several strong assumptions about the fisheries and the life history or condition of the stocks so the assessment can be performed with limited data. For instance, most catch-based models rely primarily on a time series of removals and assumptions about relative stock depletion, with limited or no explicit incorporation of life-history processes such as growth, mortality, or recruitment. Rather than reducing reliance on assumptions, several catch-based methods have been developed to formalize these assumptions explicitly within a statistical framework. Examples include DB-SRA (Dick and MacCall, 2011) and CMSY++ (Froese et al., 2023, 2021; Martell and Froese, 2013), which rely primarily on time series of removals combined with prior information on stock productivity and relative depletion. When removal data are inadequate or unavailable, methods based on length compositions may prove more appropriate, resulting in an estimate of relative spawning potential or stock status and/or the fishing mortality that led to those values. In general, structured length-based models assume that selectivity and fishing mortality derive the length composition. Examples of length-based methods include the Length-Based Spawning Potential Ratio (LBSPR; Horczyk et al., 2015c, 2015b), which is designed to estimate the remaining spawning potential of a stock following fishing activity, and the Length-based Bayesian Biomass estimator (LBB; Froese et al., 2018), which can operate using only length composition data but internally relies on assumed or informed life-history to assess stock status. Additionally, there are integrated models that can function as catch-only or length-only, as well as integrating these data. For instance, the Stock Assessment Continuum (SAC) Tool (Cope, 2024a, 2024b) was developed to address data-limited scenarios using the age-structure framework of Stock Synthesis (SS3; Cope, 2013; Methot and Wetzel, 2013). As integrated models are more flexible regarding the types of data available, a variety of data-limited methods can be built within the same modelling framework (Cope, 2024a). The SAC Tool represents an advancement over traditional fish stock assessment methods, such as the Thompson and Bell model for example, by providing a more comprehensive approach to data-limited assessments. While the Thompson and Bell model relies on static equilibrium assumptions and primarily focuses on yield-per-recruit calculations, the SAC Tool incorporates dynamic population processes within an age-structured framework. This allows for a more realistic representation of stock dynamics, including recruitment, variability and time-varying fishing pressure. Additionally, the SAC Tool's ability to integrate diverse data types within a single framework enhances its applicability across a broader range of fisheries scenarios, improving assessment robustness and management decisions.

In Brazil, especially on the northeast coast, most fisheries are considered data-limited as they are mostly characterized by its artisanal and small-scale features (IBAMA, 2007). Small-scale fishing fleets from the Northeast Brazil are mainly composed of small/medium size, which can vary from subsistence to commercial fishing, and can present a high diversity of fishing gears and exploited species (ICMBio, 2021). In the 1990s, the Brazilian government funded the REVIZEE project to explore the potential biological resources of the country's Exclusive Economic Zone (Lessa et al., 2009). In the scope of the REVIZEE project several studies were performed, including stock assessment for a variety of fish species (Lessa et al., 2009; MMA, 2006). Currently, data on these stocks, especially associated with small-scale and artisanal fisheries, are limited due to interruptions in national data collection programs over recent decades and the discontinuation of the Brazilian national fisheries monitoring system. As a result, many

693 stocks are considered data-limited, motivating the development of alternative sources of
694 information such as reconstructions of catch time series, which have become a crucial
695 input for stock assessment in data-limited contexts (Freire et al. 2021).

696 Here, a methodological case study is presented in which a stock previously
697 assessed under the REVIZEE project is used as benchmark to evaluate the behavior of a
698 data-limited assessment framework. Using the same dataset, alternative stock assessment
699 configurations using different data types within the same modelling framework were
700 applied to examine how outputs respond to reduced information and different model
701 assumptions. The *Ocyurus chrysurus* stock was selected as a case of study due to its
702 socioeconomic importance and because it has been the subject of robust assessment
703 conducted during the REVIZEE project. From this perspective, this study proposes an
704 evaluation of the performance of an integrated assessment framework under data-limited
705 conditions to assess stocks of demersal fish species from the northeastern coast of Brazil.

706 *Methods*

707 *Data collection*

708 The stock area corresponds to the Northeast Brazil Marine Ecoregion, as defined
709 by Spalding et al. (2007) that was also used in the original REVIZEE assessment (Figure
710 1). Historical landings of *Ocyurus chrysurus* were compiled from two primary sources:
711 (1) the comprehensive reconstruction of commercial landings from 1950–2015 for
712 artisanal and industrial fleets by Freire et al. (2021), and (2) the Sea Around Us catch
713 database (<https://www.seaaroundus.org/>). Both datasets were filtered to include only
714 landings within the Northeast Brazil Marine Ecoregion (ME; Spalding et al., 2007). To
715 match the temporal scope of the original REVIZEE assessment, the landings time series
716 used here was restricted to 1950–2000 (Figure 2). Length composition data originated
717 from the REVIZEE Program (MMA, 2006; Lessa et al., 2009). Samples were collected
718 between 1996 and 2000 from industrial longline fleets operating off Northeast Brazil,
719 responsible for over 95% of landings for *O. chrysurus* during the period (IBAMA, 2008).
720 Monthly length samples were obtained from landing points within the stock area (Figure
721 3). These data represent the primary source of empirical information used in the SAC
722 Tool assessment. Life-history parameters matched those used in the original assessment
723 (Frédou et al., 2009a) and are summarized in Table 1. These parameters describe key
724 biological processes, including growth, natural mortality, maturation, and reproduction,
725 and are used to parametrize the population model within the SAC Tool. Due to the lack
726 of quantitative fecundity information, individual egg production is assumed proportion to
727 weight, thus the weight-fecundity parameters were assumed equal to the weight–length
728 relationship.

729 *Base model assessment*

730 The benchmark for comparison was the original *O. chrysurus* assessment
731 conducted in the early 2000s under the REVIZEE Project (Lessa et al., 2009). This
732 assessment applied the Thompson & Bell yield-per-recruit framework (Thompson and
733 Bell, 1934) in combination with an age-based Virtual Population Analysis (Gulland,
734 1965), using catch-at-age data, natural mortality estimated from the Ault et al. (1998)
735 survivorship model, and life-history parameters consistent with an unexploited
736 equilibrium assumption. Reference points derived from the analysis were $F_{0.1}$, E_{max} ,
737 and $E_{current}$, the latter exceeding the conservative exploitation threshold ($E > 0.5$),
738 indicating overexploitation. This original assessment was treated as the base case, and its

739 published results were used directly as a base-case benchmark for the SAC Tool
740 scenarios.

741 *SAC Tool framework*

742 All data-limited scenarios were fitted using the Stock Assessment Continuum
743 Tool (Cope 2024a,b), which operates within the age-structured Stock Synthesis (SS3)
744 modelling framework (Methot and Wetzel, 2013). Required biological and fishery inputs
745 included: natural mortality (M), von Bertalanffy growth parameters (L_∞ , K , t_0), maturity
746 and selectivity ogives, weight–length parameters, Beverton-Holt stock-recruit parameters
747 (steepness $h = 0.7$; $\ln R_0$ estimated by the model), landings (when applicable), and length
748 compositions. Initial selectivity parameters were derived from a catch-curve analysis
749 (TropFishR; Mildenerger et al., 2017) and subsequently estimated by the model.
750 Reference points were stock spawning biomass at maximum sustainable yield (SSB_{msy})
751 and fishing mortality at MSY (F_{msy}). The overall workflow of the SAC Tool, including
752 data inputs, model structure, and outputs, is summarized in Figure 4.

753 *Model benchmarking strategy*

754 Before applying data-limited methods to new fisheries, model performance is
755 often evaluated through simulation testing (e.g., pseudo-data fits; Chong et al., 2020; Pons
756 et al., 2020; Punt et al., 2016; Wetzel & Punt, 2015) or by re-fitting alternative models to
757 data from a well-established assessment (Rudd et al. 2021). Here, we adopted the latter
758 approach: the data from the REVIZEE assessment were re-analyzed under six data-
759 limited configurations to evaluate whether the SAC Tool could reproduce the qualitative
760 and quantitative outcomes of the traditional, data-rich assessment. This benchmarking
761 approach enables validation of model behavior under realistic data conditions.

762 *Scenario design*

763 The scenarios were designed to represent realistic assessment configurations commonly
764 encountered in data-limited fisheries. Rather than exploring arbitrary model variants, the selected
765 scenarios reflect standard uses of the SAC Tool, allowing evaluation on how assessment outputs
766 respond to progressively reduced information. Each scenario corresponds to a plausible
767 combination of data availability and modeling assumptions routinely applied when landing
768 histories, recruitment information, or complete length time series are unavailable.

- 769 1. SC1 – Integrated (catch + lengths): Includes landings (1950–2000), length
770 compositions (1996–2000), life-history parameters, and freely estimated annual
771 recruitment.
- 772 2. SC2 – Integrated without recruitment estimation: Same as SC1, but recruitment
773 deviations constrained to zero, assuming a deterministic stock-recruit relationship.
- 774 3. SC3 – Integrated with truncated length data: Only the final three years of length
775 compositions included; landings and recruitment estimation retained.
- 776 4. SC4 – Length-only (constant catch formulation): Landings excluded; lengths drive
777 estimation of an average F over time based on an internally generated constant-catch
778 assumption.
- 779 5. SC5 – Length-only without recruitment estimation: Same as SC4 but without
780 recruitment deviations.
- 781 6. SC6 – Length-only (direct F estimation): Landings excluded; F estimated directly
782 from length compositions and M , analogous to the Length-based Integrated Mixed
783 Effects (LIME) approach of Rudd & Thorson (2018).

784 All scenarios used identical life-history inputs to isolate the influence of data
785 availability on model behavior.

786 *Performance metrics*

787 Performance was evaluated by analyzing consistency in the assessment outputs
788 across scenarios. Only a limited set of metrics can be meaningfully compared due to
789 fundamental differences between the Thompson & Bell framework and the SAC Tool.
790 Fishing mortality (F) was therefore selected as the primary quantitative metric, as is
791 estimated in both approaches. Stock status was compared qualitatively using relative
792 biomass indicators. Although depletion metrics such as SSB/SSB_{msy} can, in principle, be
793 derived under both frameworks, the historical Thompson & Bell assessment did not report
794 SSB_{msy} explicitly, and stable SSB_{msy} estimates were not available across all data-limited
795 SAC Tool scenarios. As a result, direct numerical comparisons of biomass-based status
796 metrics were not possible. Instead, SAC Tool outputs were interpreted using its internal
797 limit reference point ($0.25 \times SSB_0$) as a screening threshold for qualitative stock status
798 classification.

799 *Results*

800 Variability among scenarios was evaluated descriptively based on the magnitude
801 and spread of model outputs. Across all configurations, scenarios converged qualitatively
802 to the same stock-status conclusion. As expected for data-limited assessments, reductions
803 in available information were associated with wider uncertainty intervals, whereas
804 scenarios incorporating more data exhibited narrower confidence intervals, indicating
805 greater apparent precision. The reference scenario presented a total initial biomass (B_0)
806 of 11,944 T, an F of 0.274, relative fishing mortality of 4.98, indicating fishing pressure
807 was far above the sustainable limit, and the exploitation rate of 0.617 indicated the stock
808 was overexploited (Frédou et al., 2009a; Lessa et al., 2009).

809 Across all scenarios, results consistently indicated an overfished stock and
810 highlighted elevated fishing mortality (Table 2). The SAC Tool also reproduced the main
811 features of the observed length compositions, with differences in fit reflecting expected
812 reductions in information content among length-only configurations (Figure 5).

813 *Integrated model scenarios*

814 The three first scenarios were set as integrated models, as it considered landings
815 and length compositions as inputs. In these cases, scale and status of the stock were
816 estimable as the models have the minimum data required to do so.

817 Scenario 1 estimated total spawning biomass (B_0) of 18,032 tons, with the current
818 spawning biomass (SSB) at only around 13% of B_0 (Figure 6a) well below the SAC Tool
819 internal limit reference point ($0.25 \times SSB_0$). In addition to the biomass being well below
820 desired levels, the exploitation ratio F/F_{msy} of 2.52 shows that contemporary fishing
821 mortality ($F = 0.35$) had been considerably higher than the level associated with MSY
822 (Figure 7a). Overall, SC1 depicts a stock still well below historical productivity and likely
823 experiencing continued overexploitation.

824 Scenario 2 reported B_0 of 18,767 tons and gave the most relatively optimistic
825 outcome among the integrated models, with SSB representing 22% of B_0 (Figure 6b).
826 Fishing mortality remains above optimal levels ($F = 0.22$; $F/F_{msy} = 1.60$), but much lower
827 than in SC1 and SC3 (Figure 7b). The absence of recruitment estimates in SC2 resulted
828 in slightly lower apparent fishing pressure and higher relative biomass, as the model
829 compensated for constant recruitment by increasing the estimated population scale ($\ln R_0$),

830 thus accommodating the same landings time series with a lower estimated F , highlighting
831 the role of recruitment dynamics in perceived stock depletion in integrated models.

832 The third scenario presented B_0 of 18,022 tons and showed a more depleted
833 condition, with SSB at 12% of B_0 and only 48% of SSB_{msy} (Figure 6c). These values
834 suggest that current biomass is substantially below both historical and MSY -based
835 reference points. Fishing pressure remains high ($F = 0.36$; $F/F_{msy} = 3.32$), consistent with
836 ongoing overfishing (Figure 7c). Although limiting the length data to the last three years,
837 $SC3$ still detects high exploitation and low relative biomass, showing that recent length
838 information can capture the current overfished state when recruitment estimates are
839 included.

840 All integrated scenarios ($SC1$ – $SC3$) produced relative biomass estimates
841 ($SSB/SSB_0 = 0.12 - 0.22$) that consistently indicated an overexploited stock, in agreement
842 with the status inferred from the base assessment. Although the point estimates differed
843 slightly, all values remained well below unfished levels, confirming substantial depletion.
844 Fishing mortality estimates were also consistent across integrated scenarios ($1.6 - 2.58 \times$
845 F_{msy}), indicating that exploitation remains above the maximum sustainable level. When
846 compared to the base model $F_{current} = 0.274$, the integrated scenarios suggest similar or
847 higher fishing pressure, with $SC2$ producing the lowest estimate, possibly due to the
848 removal of recruitment effects. Overall, the integrated models behaved coherently with
849 the base assessment, reinforcing that the stock remains overexploited.

850 *Length-only scenarios*

851 The last three scenarios were composed of length-only models, where the time
852 series of landings were removed from the analysis. In these scenarios stock scale is not
853 estimated, only stock status and fishing pressure.

854 Using a length-only, constant-catch approach, $SC4$ presented $F = 0.51$ (Figure 7d)
855 and $SSB/SSB_0 = 0.15$ (Figure 6d). This scenario tends to overestimate fishing mortality
856 compared to integrated models, reflecting the model's emphasis on fitting length
857 composition patterns under the assumption of relatively constant catch.

858 Excluding recruitment in $SC5$ led to $F = 0.39$ (Figure 7e) and $SSB/SSB_0 = 0.21$
859 (Figure 6e). Relative to $SC4$, this reduces the apparent fishing pressure and increases
860 relative biomass, mirroring the effect of recruitment removal observed in integrated
861 models. This reflects the same pattern observed in the integrated scenarios, where
862 removing recruitment variability leads to higher biomass estimates and lower apparent
863 fishing pressure.

864 $SC6$, which estimates F directly from length data, presented $F = 0.43$ (Figure 7f)
865 and $SSB/SSB_0 = 0.17$ (Figure 6f). While the point estimate of F is closer to integrated
866 model values, the uncertainty around it is extremely large. This indicates that length-only
867 estimate- F approaches may produce plausible point estimates but large uncertainty
868 intervals, highlighting the sensitivity of these models to data limitations and assumptions,
869 though realistic in the expectations that less data should generally lead to more
870 uncertainty.

871 Despite differences in magnitude and precision, all length-only scenarios
872 consistently indicated overexploitation ($SSB/SSB_0 < 0.25$) and fishing mortality above
873 MSY levels ($F/F_{msy} > 1$). However, their uncertainty patterns varied notably: $SC4$
874 produced the highest F and lowest biomass, $SC5$ moderated these values when
875 recruitment was fixed, and $SC6$ presented plausible point estimates but large uncertainty
876 intervals driven by parametrization.

877 *Scenario Synthesis*

878 The ensemble of model scenarios and the historical base assessment converge on
879 a single management-relevant conclusion: *Ocyurus chrysurus* in the study area is
880 currently overfished and subject to overfishing (Figure 8). Integrated assessments (SC1 –
881 SC3) produced F medians of 0.22 - 0.36 ($F/F_{msy} = 1.6 - 2.6$) with $SSB/SSB_0 = 0.12 - 0.22$.
882 Length-only approaches (SC4 – SC6) were more sensitive to modeling assumptions, SC4
883 returned the highest F (0.51; $F/F_{msy} = 3.8$), SC5 decreased F when recruitment was
884 removed, and SC6 yielded a point estimate comparable to integrated models but produced
885 unrealistic uncertainty that precludes confident management inference from its intervals.
886 These patterns are consistent with the conclusions of the original Thompson-Bell
887 assessment conducted under the REVIZEE Project, which indicated exploitation rates
888 above sustainable levels. Despite methodological and reference points differences
889 between models, both indicated that exploitation was exceeding sustainable levels and
890 rebuilding measures were required.

891 *Discussion*

892 It is expected that assessment methods will respond differently to reductions in
893 data availability, leading to contrasts in uncertainty and inferred stock status. In general,
894 when fewer data are available, uncertainty increases, a desirable property for data-limited
895 approaches (Cope, 2024a). The comparative performance of the models revealed clear
896 contrasts in biomass and fishing mortality estimates, offering insight into the suitability
897 of different assessment frameworks for demersal fisheries in Northeast Brazil. The
898 assessment conducted for *O. chrysurus* in the early 2000s, one of the first for this and
899 other lutjanids (Frédou et al., 2009a; MMA, 2006), benefited from relatively rich datasets,
900 including age- and length-composition information and life-history parameters. This
901 assessment served as the base model and benchmark for evaluating all other scenarios.
902 Although some differences across scenarios are expected due to varying data inputs,
903 qualitative consistency among outputs is generally anticipated.

904 *Assessing performance*

905 The comparative assessment across the six data-limited scenarios and the base
906 model revealed consistent signals of stock depletion for *O. chrysurus*, despite differences
907 in data richness, modelling assumptions, and analytical frameworks. Several scenarios
908 excluded landing data, others constrained recruitment variability, and the relative
909 weighting of biological or fishery information. Nevertheless, all scenarios consistently
910 classified the stock as overfished, mirroring the results of the early 2000's assessment.
911 Although these approaches rely on fundamentally different modelling philosophy, the
912 convergence of outcomes suggest that the available data, even when simplified, can carry
913 a strong and non-conflicting signal of unsustainable exploitation. This consistency across
914 methods may indicate that the overexploited status is robust to violations of some
915 biological assumptions and to alternative ways of weighting data sources within the
916 assessment.

917 The scenarios were categorized into two main groups: the integrated models (SC1
918 – SC3) and length-only models (SC4 – SC6). Integrated scenarios, which incorporated
919 landings, lengths, and life-history parameters, produced the most coherent and stable
920 estimates of F relative to the base case. These models rely on removals to approximate
921 stock scale and use lengths to inform relative stock status and selectivity, leading to more
922 balanced trade-offs between information richness and model assumptions (Cope 2024a;
923 Rudd et al. 2021). Among these, SC2 produced the most precise stock-status estimates,
924 driven by the effect of constraining recruitment deviations. When recruitment is

925 deterministic, variability in length structure cannot be transmitted into biomass dynamics,
926 reducing uncertainty intervals and biased relative biomass estimates (Rudd and Thorson,
927 2018, Thorson et al, 2019). This pattern matches expectations: constraining recruitment
928 variability increases statistical precision but may underrepresent the true uncertainty
929 associated with stock dynamics, particularly for a moderately long-lived tropical
930 demersal species, such as *O. chrysurus*. Length-only scenarios provide estimates with
931 larger variance, as expected and desirable for models relying solely on size composition
932 data. Without historical landings, these models cannot resolve absolute scale and instead
933 interpret size information through assumptions about mortality, growth, and selectivity.
934 Uncertainty increased substantially in scenarios where F is estimated directly from length
935 data, a byproduct of using asymptotic variability; Bayesian variance estimation could get
936 around this issue (Punt and Hilborn, 1997).

937 Nonetheless, all length-only scenarios identified the stock as depleted relative to
938 sustainable levels, matching the F estimates of those reported in the historical base
939 assessment. Direct quantitative comparison between the historical Thompson-Bell
940 assessment and the SAC Tool outputs is inherently limited by the distinct conceptual
941 frameworks of each approach. The earlier assessment relied on yield-per-recruit and
942 exploitation-per-recruit reference points (e.g., $F_{0.1}$, E_{max} , $E_{current}$), whereas the SAC Tool
943 produces MSY-based estimates, such as F_{MSY} and relative spawning biomass. Because
944 these metrics are not directly compatible, fishing mortality (F) represents the only
945 parameter that can be meaningfully compared across frameworks. Even so, F alone does
946 not fully define stock status, which ultimately depends on the resulting spawning stock
947 biomass and recruitment dynamics. The close alignment of F estimates, despite
948 differences in data richness, model structure, and uncertainty, demonstrates that the SAC
949 Tool and the historical Thompson-Bell assessment can infer terminal-year fishing
950 pressure from population structure interpreted through species-specific life-history
951 constraints, supporting the argument that data-limited frameworks, when properly
952 specified, can produce management-relevant indicators that agree with established
953 assessments even in data-poor contexts.

954 Overall, despite differences in model structure and precision, every scenario
955 agreed with the base scenario on the fundamental stock-status classification: *O. chrysurus*
956 is overfished and undergoing overfishing. This consistency across disparate modelling
957 approaches strengthens confidence in the qualitative conclusion, even when quantitative
958 outputs differ or carry substantial uncertainty. The more flexible modelling framework of
959 SS3 allows more options for future data inclusion and model specification.

960 *Recruitment estimates in the SAC Tool and implications*

961 Recruitment estimation in the SAC Tool is grounded in the Stock Synthesis
962 framework, which uses deviations from the Beverton and Holt stock-recruit relationship
963 (Cope 2013; Methot and Wetzel, 2013). In data-limited configurations, recruitment
964 deviations are informed indirectly through the interaction of length composition, growth,
965 natural mortality, and selectivity patterns (Cope 2024a). Lengths act as proxies for age
966 structure, enabling the model to infer periods of stronger or weaker recruitment based on
967 shifts in size frequency, though with substantial sensitivity to the assumed or estimated
968 selectivity curve (Rudd and Thorson, 2018). This approach provides a valuable
969 advantage: it avoids imposing the strict assumption of constant recruitment, which often
970 underlies simpler length-based methods, such as LBSPR (Hordyk et al. 2015a). By
971 allowing empirical deviations, the SAC Tool incorporates biologically plausible
972 variability and can better resolve the separation between fishing mortality and recruitment
973 dynamics, two processes that otherwise become confounded in purely length-based

974 perspectives (Cope, 2024a). However, this flexibility comes with trade-offs. When
975 recruitment is allowed to vary (e.g. SC1 and SC3), uncertainty naturally increases, and
976 this effect persists even in data-rich scenarios. Patterns in length data are not always
977 strong enough to uniquely identify recruitment pulses, particularly in tropical
978 multispecies fisheries where selectivity is broad and growth rates are rapid (Chong et al.,
979 2020). This uncertainty should not be interpreted as a flaw but rather as a desirable
980 reflection of data-limited assessments, preventing the artificial precision that emerges
981 when structural variability is suppressed (Cope 2024a; Rudd et al. 2021). Conversely,
982 when recruitment is fixed (e.g. SC2 and SC5), the model produces narrower confidence
983 intervals and smoother biomass trajectories but at the cost of biological realism. These
984 scenarios tend to overestimate relative biomass and underestimate fishing mortality
985 because they cannot attribute observed size-structure compression to poor recruitment. In
986 practical stock-assessment contexts, this reinforces the need for caution when disabling
987 recruitment estimation: it may stabilize outputs but risks bias toward optimistic status
988 classifications.

989 *“Constant-catch” vs “F-estimate” in length-only models*

990 The two length-only configurations in the SAC Tool, constant catch (SC4, SC5)
991 and direct F estimation (SC6) represent distinct conceptual approaches to interpreting size
992 compositions in the absence of removals. The “constant catch” approach uses an assumed
993 or approximate average catch to anchor the scale of fishing pressure. In this formulation,
994 F is derived similarly to LBSPR and related length-based mortality estimators (Hordyk et
995 al. 2015c; Cope 2024a): the model fits a catch-curve like mortality signal to the observed
996 size distribution, with the assumed catch level constraining the scale of depletion. When
997 there are multiple years of length compositions, it fits them as a proper time series without
998 the need to assume an equilibrium fishing mortality. This method generally produces
999 more stable estimates because fishing mortality is translated through the assumed
1000 constant catch (Cope, 2024a). Thus, uncertainty is typically lower than in direct F-
1001 estimation models. By contrast, the F-estimate approach (SC6) attempts to infer fishing
1002 mortality directly from length data and natural mortality alone, closely resembling the
1003 conceptual logic of LIME (Rudd & Thorson 2018). Without any catch information, the
1004 model must rely exclusively on the changes in length-frequency assuming a constant, but
1005 estimated, selectivity, and the assumed growth parameters. This results in far greater
1006 uncertainty and instability, especially when length samples are sparse in some years or
1007 recruitment variability is high (Hordyk et al, 2015b; Rudd and Thorson, 2018). The wide
1008 and occasionally unrealistic confidence intervals in SC6 are consistent with expectations
1009 for models that attempt to solve simultaneously for selectivity, recruitment, and fishing
1010 mortality from a single dataset with limited samples.

1011 From a methodological perspective, the constant catch formulation stabilizes
1012 fishing mortality estimation by imposing an artificial, fixed catch to guide the model,
1013 reducing confounding between F, recruitment, and selectivity, but assumes a fixed catch
1014 history. This approach generally produces more precise estimates and narrower
1015 confidence intervals, but its reliance on an assumed constant catch means that absolute F
1016 values may be biased if the true fishing pattern differs from the assumption, and temporal
1017 variability in fishing effort cannot be captured. In contrast, the direct F-estimation
1018 formulation infers fishing mortality solely from length data and natural mortality, without
1019 assuming any catch. While this allows the model to operate in a real context of “length-
1020 only”, it results in far greater uncertainty, and higher sensitivity to sparse length data,
1021 recruitment variability, and mis-specified selectivity. Both approaches are reliable within
1022 their respective contexts, and the SAC Tool’s flexibility allows analysts to select the

1023 formulation most suitable for the available data while acknowledging the inherent trade-
1024 offs (i.e., each method has its pros and cons).

1025 *Management implications*

1026 The consistent conclusion across all SAC Tool scenarios that *O. chrysurus* is
1027 overfished and undergoing overfishing closely matches the base scenario, which
1028 represents the original assessment from the early 2000s and serves as the reference model.
1029 Importantly, this alignment occurs despite the testing of different assumptions that were
1030 implicit in the earlier Thompson–Bell and VPA framework. Scenarios assuming constant
1031 recruitment, variable recruitment, or constant catch represent alternative representations
1032 of population dynamics and data availability that were not explored in the historical
1033 assessment. Even under these simplified configurations, all scenarios detected depletion,
1034 indicating that the signal of overexploitation is robust to different structural assumptions
1035 about population dynamics and data inputs. This demonstrates that contemporary data-
1036 limited methods not only reproduce key outcomes of a historically data-rich assessment
1037 but also provide a flexible framework to explicitly evaluate how these assumptions
1038 influence the inference of stock status. For multispecies demersal fisheries in Northeast
1039 Brazil, where historical landing series are incomplete and monitoring programs remain
1040 resource-limited (Frédou et al 2017; Silva et al 2021; Silva et al 2025), length-based and
1041 integrated data-limited approaches provide feasible and informative alternatives. Length
1042 samples are generally cheaper to collect, less susceptible to systemic reporting biases, and
1043 more easily validated than catch statistics (Thorson et al. 2019), making them practical
1044 first-line tools for routine monitoring in small-scale fisheries where institutional capacity
1045 is constrained (Chong et al. 2020; Medeiros-Leal et al. 2023). Nonetheless, data-limited
1046 assessments should be interpreted as components of an adaptive management system
1047 rather than definitive endpoints. As emphasized by Dowling et al. (2019, 2023),
1048 indiscriminate or uncritical application of data-limited methods can lead to
1049 misinterpretation. Transparent communication of uncertainty, continued data collection,
1050 and progressive integration of new information are therefore essential. Within this
1051 context, the SAC Tool’s flexible age-structured framework allows assessments to evolve
1052 as additional data become available, progressively strengthening inference and improving
1053 comparability with data-rich benchmarks such as the original assessment, while also
1054 clarifying how different data inputs influence model outputs and whether conflicting
1055 signals are present (Cope, 2024a).

1056 Taken together, these findings indicate that data-limited assessments may reliably
1057 replicate qualitative stock status trends observed in historical, data-rich assessments, and
1058 executed within a common framework. The consistent pattern of overfishing across all
1059 model scenarios, and their agreement with the base scenario underscores the usefulness
1060 of these methods and supports their application in contemporary management. This
1061 demonstrates that data-limited approaches are not only practical but also are also of
1062 defensible consideration for informing precautionary management, including effort
1063 reductions, harvest controls, and spatial protections, in tropical demersal fisheries where
1064 comprehensive historical data are lacking, while providing a framework that can be added
1065 to in the future as more data and life history information becomes available.

1066 *Conclusion*

1067 This study reassessed the early-2000s stock evaluation of *Ocyurus chrysurus*
1068 using contemporary data-limited approaches implemented through the SAC Tool. By
1069 benchmarking six alternative scenarios against the original data-rich assessment, we

1070 demonstrated that modern data-limited methods can consistently reproduce the same
1071 qualitative conclusion: the stock is overfished and experiencing overfishing. This
1072 alignment is notable given the substantial reduction in data inputs relative to the original
1073 assessment, and it provides strong evidence that simplified, length-based and integrated
1074 data-limited approaches are scientifically defensible for tropical demersal fisheries. The
1075 integrated scenarios, even under limited data availability, produced estimates of fishing
1076 mortality and stock depletion that closely matched the base scenario. Length-only
1077 scenarios also converged on the same qualitative stock status, despite exhibiting greater
1078 uncertainty. Collectively, these results validate the use of SAC Tool configurations across
1079 a wide data-availability gradient and highlight their utility for fisheries where long-term
1080 monitoring is incomplete or inconsistent. The successful replication of the original
1081 assessment's conclusions underscores the robustness of both the historical evaluation and
1082 the contemporary methods used here. As Brazilian fisheries face persistent gaps in data
1083 collection, these findings demonstrate that credible management advice can be generated
1084 without the extensive datasets historically required. Data-limited approaches should be
1085 viewed as an entry point into an adaptive management framework, with assessments
1086 gaining precision as new data, such as improved catch reporting, length samples, age
1087 structures, or abundance indices, are incorporated into the modeling framework. Overall,
1088 this reassessment shows that data-limited methods, when applied carefully and
1089 transparently, can play a central role in the sustainable management of tropical demersal
1090 resources. Their demonstrated ability to replicate the key conclusions of a traditional data-
1091 rich assessment provides a solid foundation for their broader adoption in Brazilian
1092 fisheries science and management.

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 1258

1259 *Tables*

1260 Table 1: Input parameters and their treatments in the Stock Synthesis – DL reference model for the mutton
 1261 snapper *Lutjanus analis* from the Northeast Brazil. Sources: 1 – Lessa, Bezerra Júnior and Nóbrega 2004; 2 – Araújo,
 1262 Martins and Costa 2002.

Parameter	Estimates	Treatment	Source
Mean age at 50% maturity A_{50}	2.60	Fixed	1
VBGF Asymptotic length L_{∞}	56.71	Fixed	2
VBGF growth coefficient k_{yr}^{-1}	0.130	Fixed	2
VBGF Age at length 0 t_0	-0.773	Fixed	2
Mean length at 50% maturity L_{50}	20.15	Fixed	1
Mean length at 95% maturity L_{95}	26.03	Fixed	1
Natural Mortality M	0.17	Fixed	1
WL relationship - α	0.035	Fixed	1
WL relationship - β	2.74	Fixed	1
Weigh - based fecundity coefficient	0.035	Fixed	-
Weigh - based fecundity exponent	2.74	Fixed	-
Steepness h	0.70	Fixed	-

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Table 2: Summarized results of the Stock Assessment Continuum Tool analysis for six assessment scenarios of *Ocyurus chrysurus* from Northeast Brazil. Scenarios represent different levels of data availability and modeling assumptions, ranging from integrated catch–length configurations to length-only approaches. The historical Thompson-Bell assessment (REVIZEE) is included as a benchmark for fishing mortality. Biomass-based depletion metrics are not available for the base model

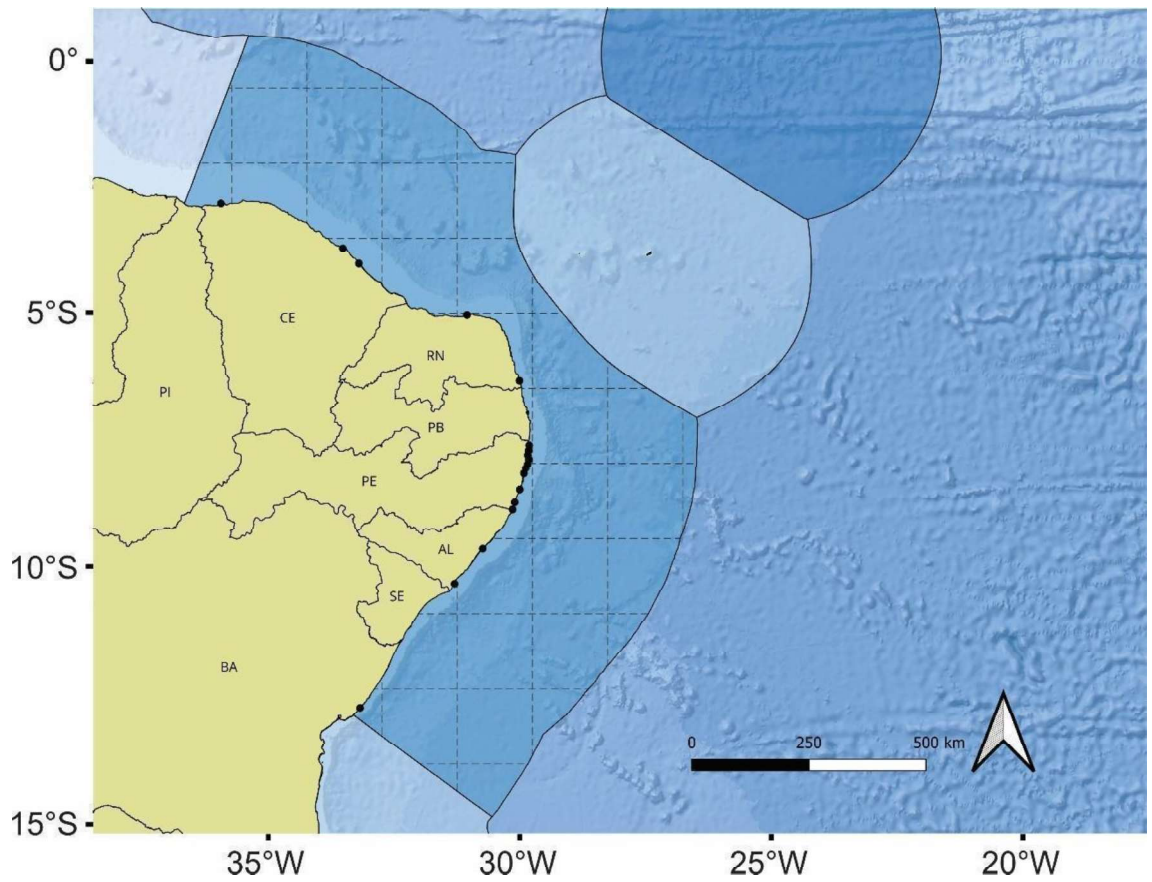
Scenario	Initial biomass (B₀)	Relative biomass (SSB/SSB₀)	Fishing mortality (F)	Relative F (F/F_{msy})	Stock status
Base model (REVIZEE)	-	-	0.274	-	Overexploited
Scenario 1	18,032 tons	0.127	0.347	2.53	Overexploited
Scenario 2	18,767 tons	0.220	0.220	1.58	Overexploited
Scenario 3	18,022 tons	0.119	0.360	3.32	Overexploited
Scenario 4	-	0.151	0.514	3.83	Overexploited
Scenario 5	-	0.210	0.389	2.87	Overexploited
Scenario 6	-	0.172	0.435	1.44	Overexploited

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Figures



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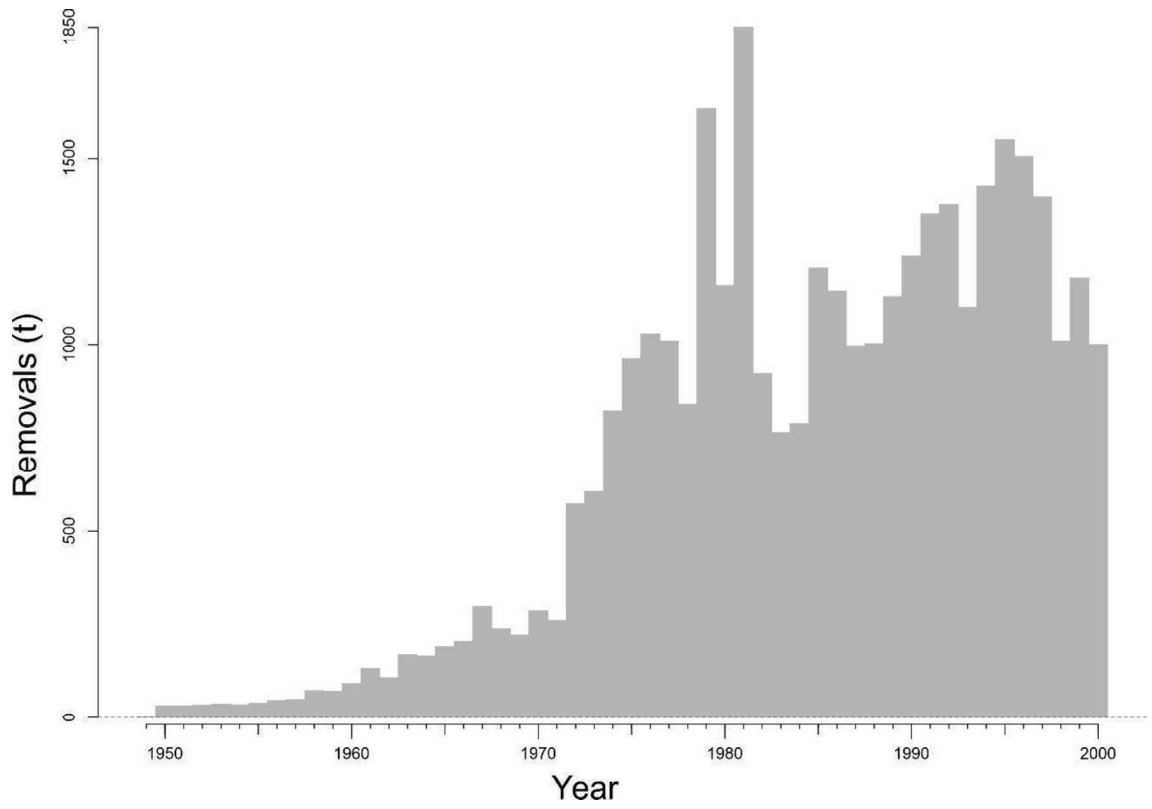
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Figure 1: Map showing the spatial domain of the *Ocyurus chrysurus* stock evaluated in this study. Black points indicate sampling or landing sites for length data. The shaded region corresponds to the Northeast Brazil Marine Ecoregion as defined by Spalding et al.

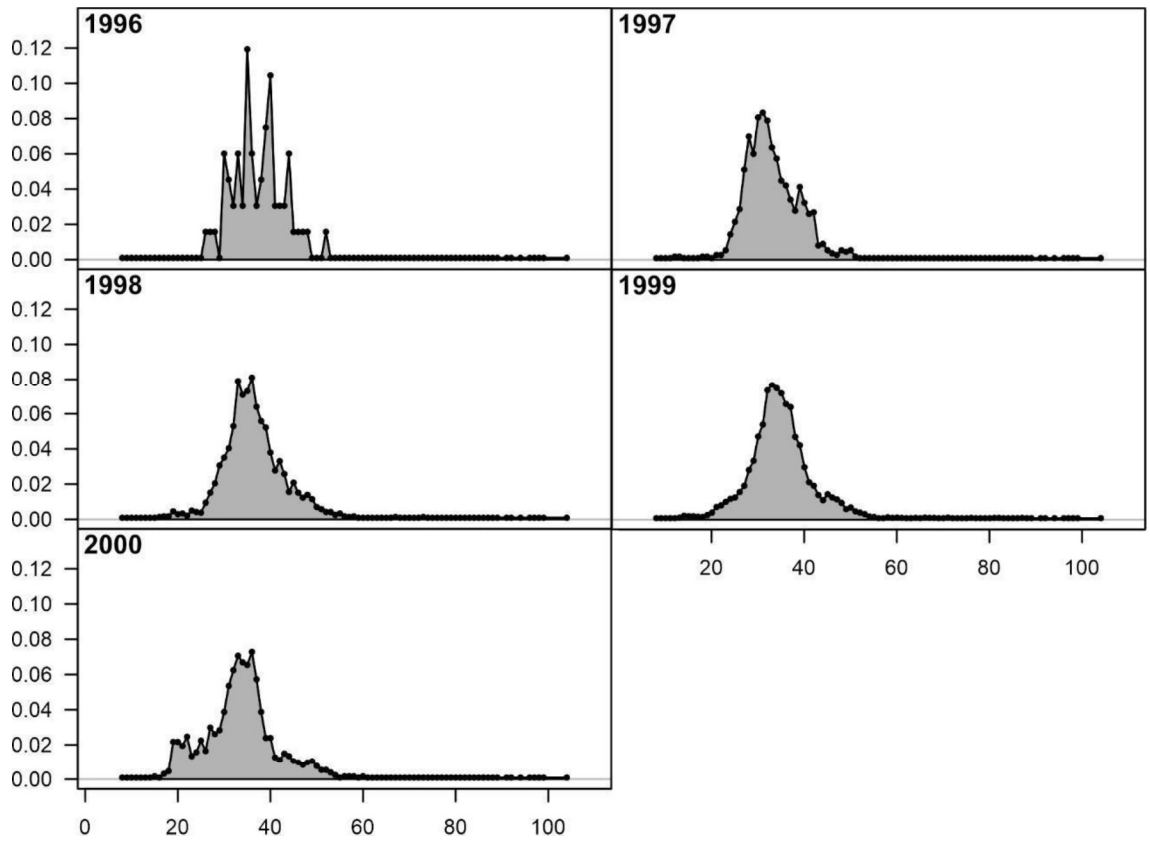
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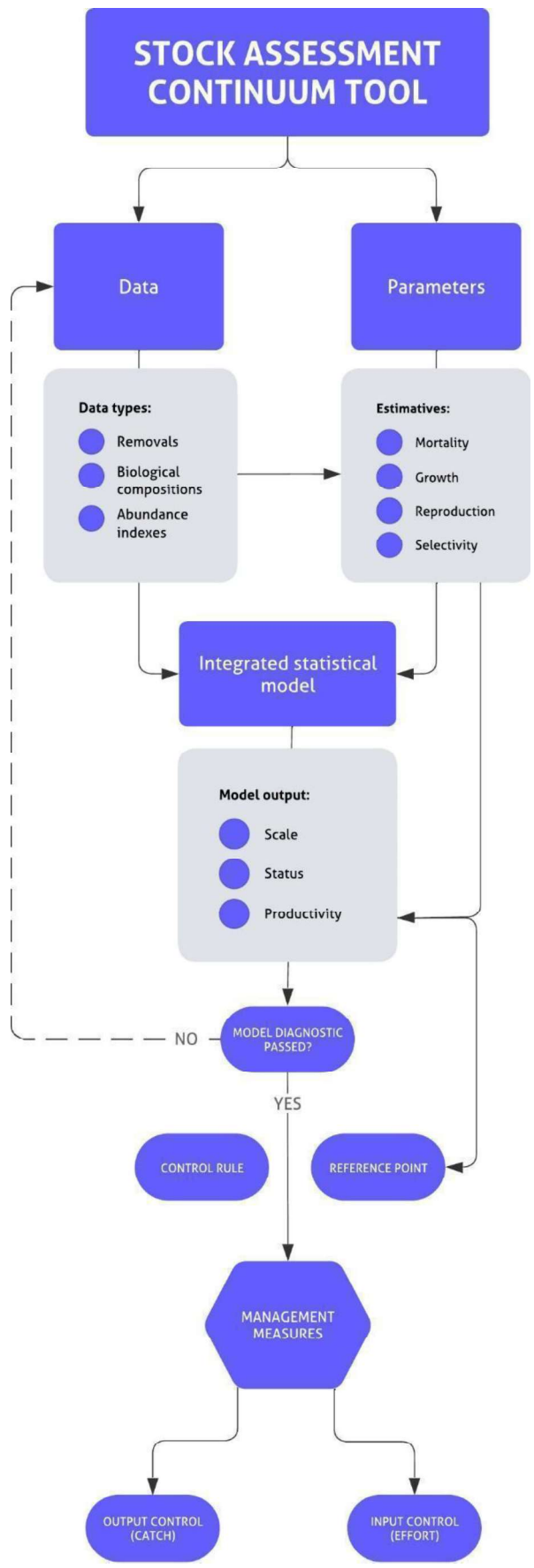
Figure 2: Time series of annual landings used as input for the integrated SAC Tool scenarios (SC1–SC3). These landings provide the stock-scale information required by the model and represent the historical exploitation of *Ocyurus chrysurus*.

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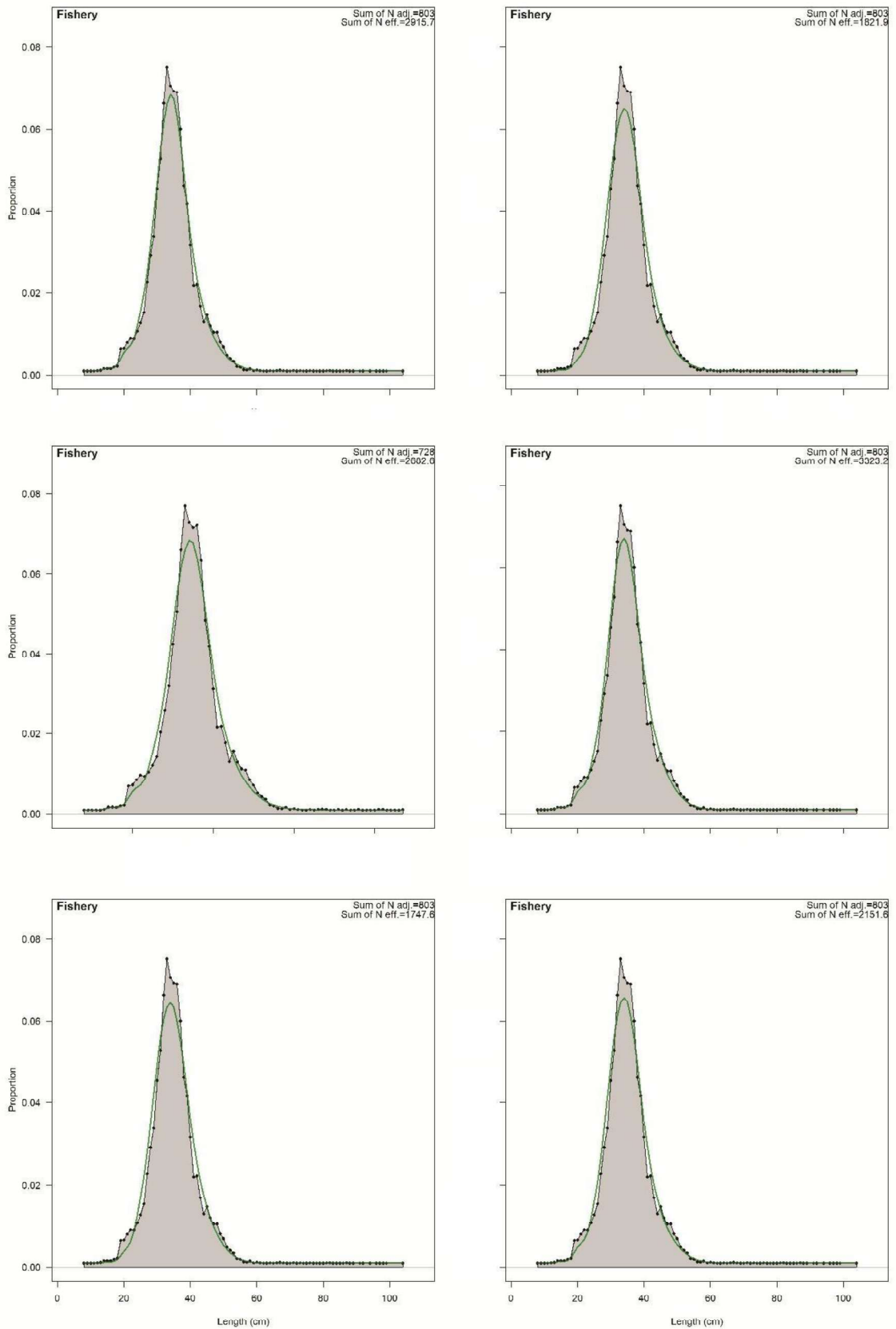
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Figure 3: Normalized length compositions used as population structure inputs for both integrated and length-only scenarios. Each panel represents one year of observations.



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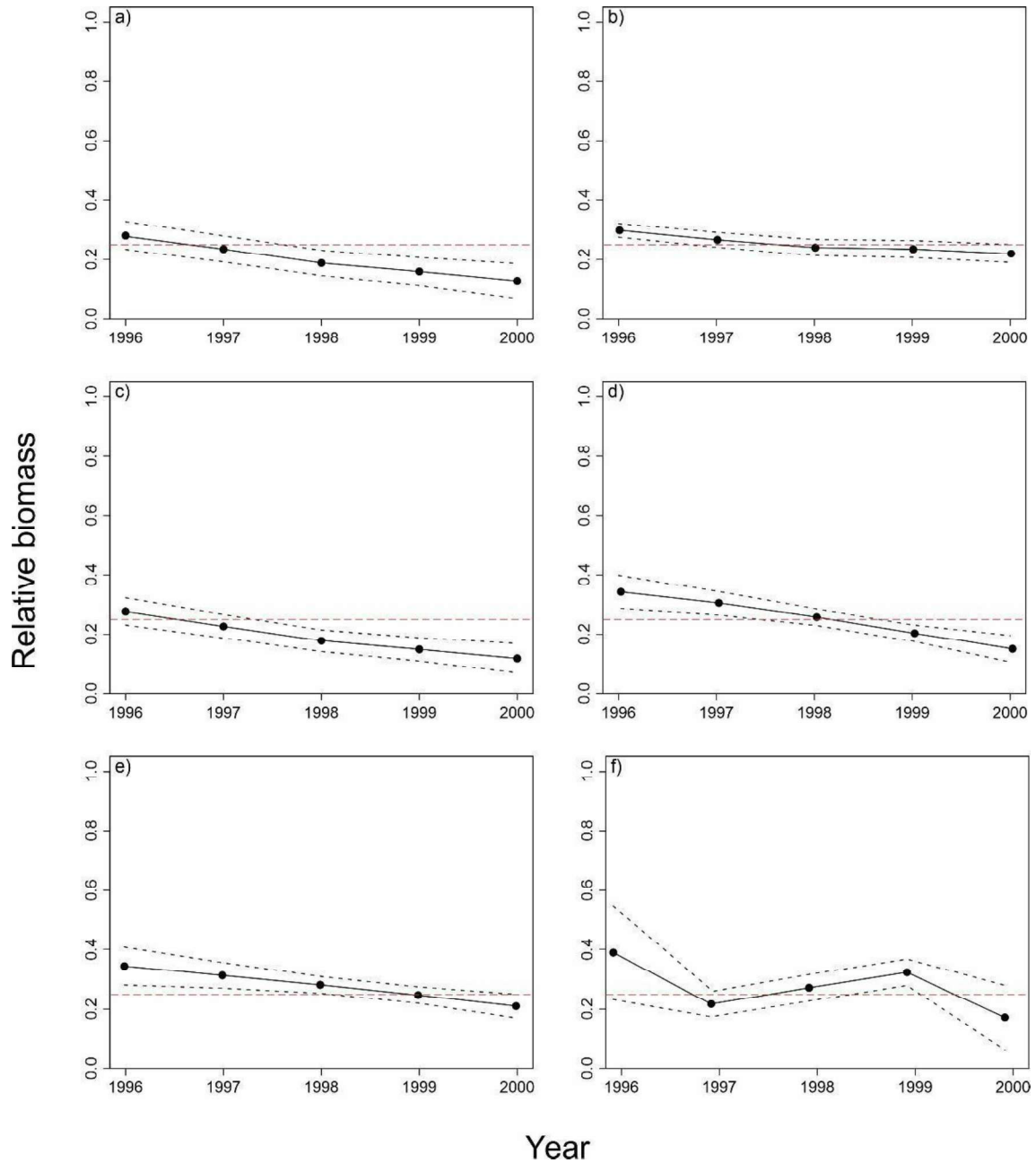
Figure 4: Diagram summarizing the SAC Tool structure, including data inputs (landings, biological compositions, abundance indices), parameter specifications (growth, mortality, selectivity, recruitment), the integrated statistical modeling framework, and derived outputs on stock scale, status, and productivity.



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1309 Figure 5: Composite fits to observed length compositions for each SAC Tool scenario. Panels (a–f)
 1310 correspond to scenarios SC1–SC6, respectively. Black points represent observed proportions, and green lines represent
 1311 model-predicted length compositions. Panels are ordered according to scenario number.

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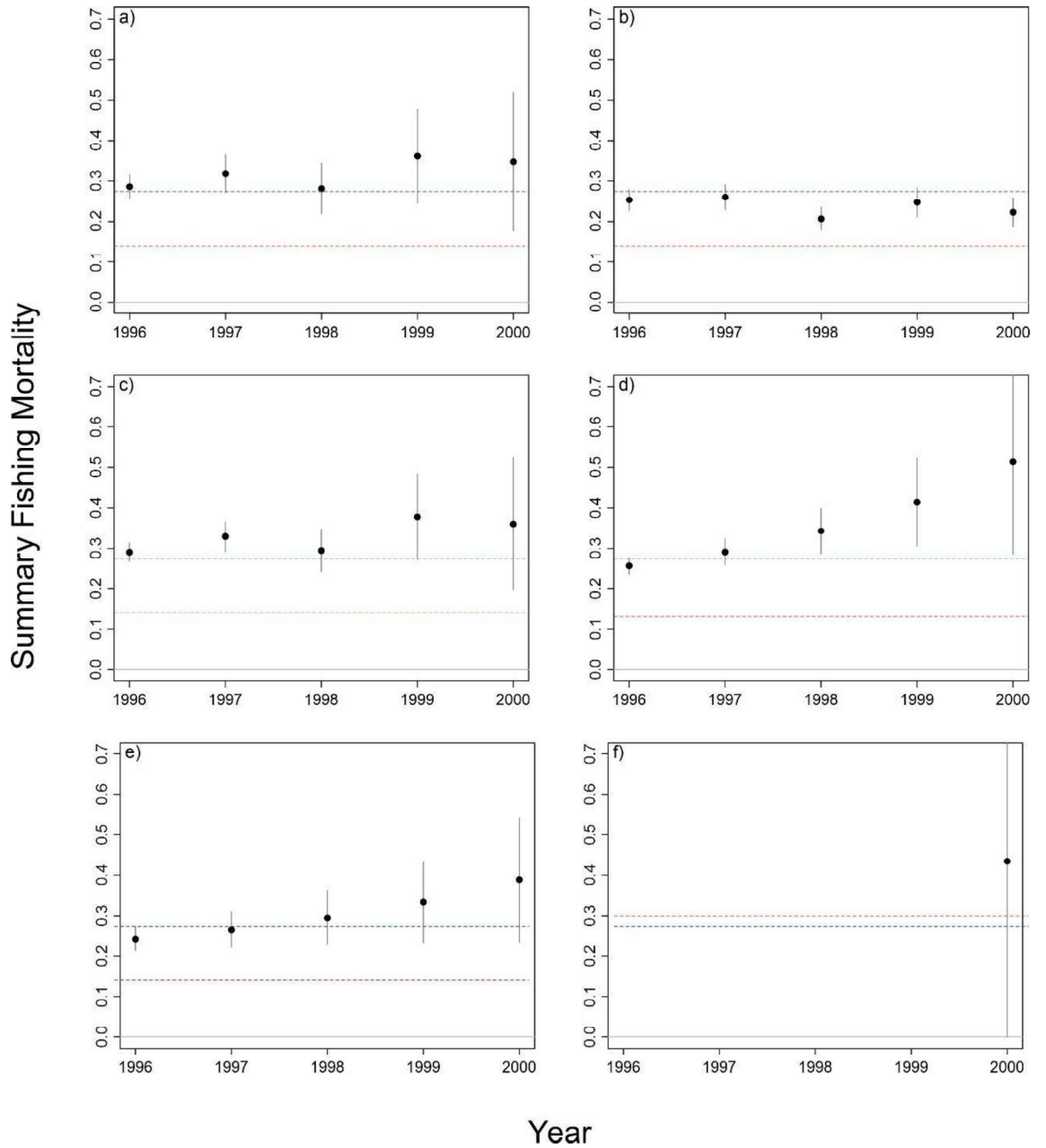


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1314 Figure 6: Estimated relative spawning biomass (SSB/SSB_0) from 1996 to 2000 for each SAC Tool scenario.
 1315 Panels (a–f) correspond to scenarios SC1–SC6, respectively. Solid black lines and dots represent the model estimate
 1316 and dashed black lines represent uncertainty intervals. Dashed red line represents the SAC Tool limit reference point
 1317 of $0.25 \times SSB_0$, used for qualitatively stock status classification.

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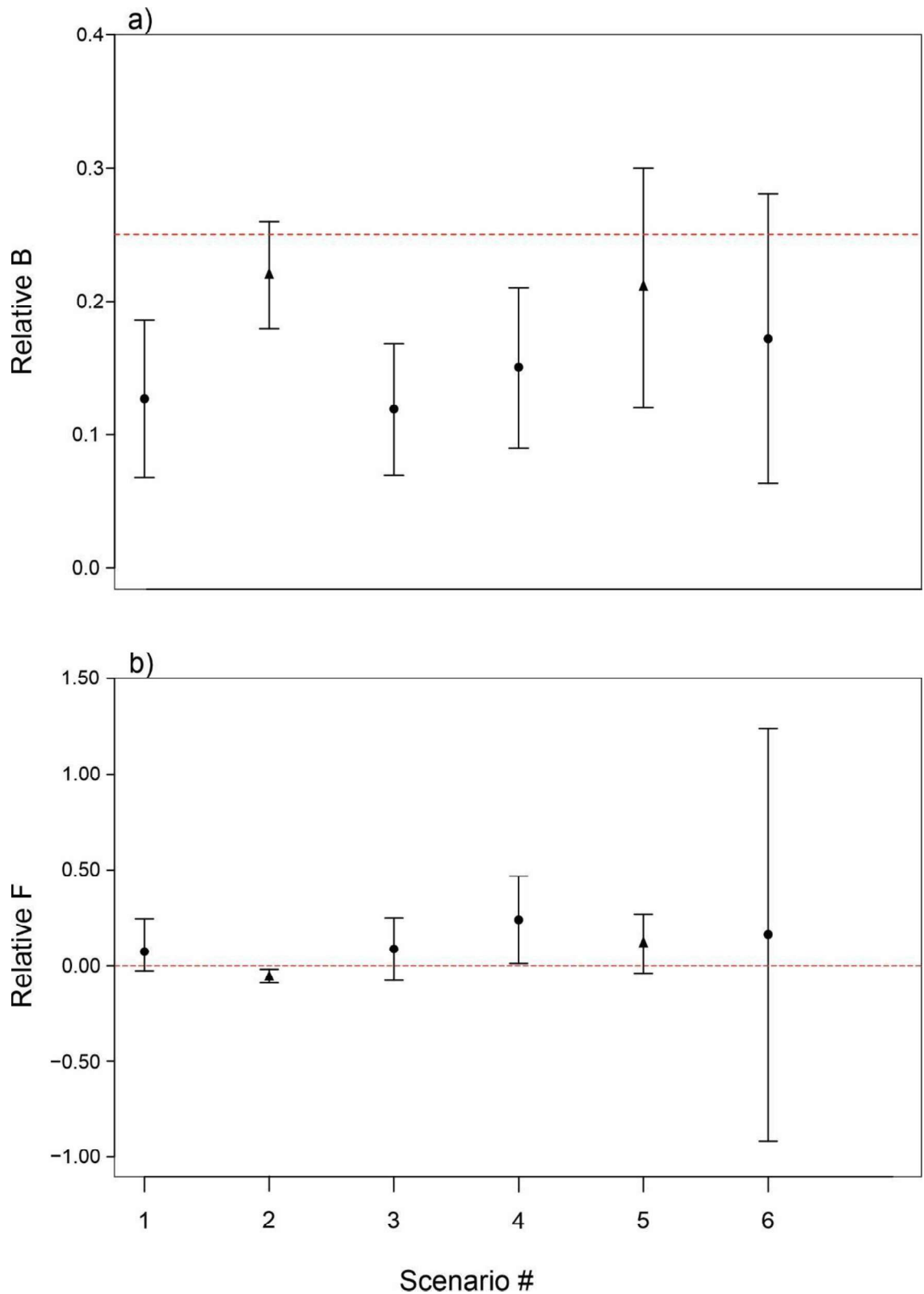
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Figure 7: Estimated fishing mortality for each SAC Tool scenario. Panels (a–f) correspond to scenarios SC1–SC6, respectively. Points represent median estimates and vertical bars indicate uncertainty intervals. The red dashed line marks the MSY-based reference point ($F/F_{msy} = 1$). The blue dotted line represents the fishing mortality estimated in the historical Thompson & Bell assessment (REVIZEE), shown for reference.



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Figure 8: Terminal-year stock status across SAC Tool scenarios: (a) relative biomass (SSB/SSB_0) for each scenario (SC1–SC6). Points show median estimates with uncertainty bars. The red dashed line marks the SAC Tool limit reference point ($0.25 B/B_0$), below which the stock is considered overfished and (b) relative fishing mortality expressed as the difference between each scenario's F estimate and the F obtained in the historical REVIZEE assessment. Values above zero indicate higher fishing mortality than the base model; values below zero indicate lower estimates. The red

1332 dashed line at zero represents equality with the base-model F. Circles indicate scenarios with estimated recruitment,
1333 whereas triangles indicate scenarios with fixed recruitment.

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Capítulo II

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ARTICLE OPEN ACCESS

20 Years Later: Updated Stock Assessment of Snappers in Northeast Brazil Using an Integrated Stock Assessment Framework

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Keywords: data limited stock assessment | fish management | groundfish | Lutjanidae | marine ecoregions | stock synthesis

ABSTRACT

We updated the stock status of *Lutjanus analis*, *Lutjanus joca*, *Lutjanus synagris*, and *Ocyurus chrysurus* harvested along the Brazilian northeastern coast. Stock boundaries were defined according to the Marine Ecoregion classifications at a finer scale, to reflect the population structure of each species. Data were exclusively from the handline fishing fleet, with removals obtained from fisheries national statistics and published commercial fishery records. Length compositions were from commercial and scientific surveys. Abundance indices were estimated from data from three projects and Brazilian official fishery statistics reports. Growth parameters were from the literature, and natural mortality was estimated using the Natural Mortality Tool. Data were analyzed using the Stock Assessment Continuum Tool. Three of the four species were currently overexploited, with populations having declined by > 70% from 1950 to 2021. In contrast, *O. chrysurus* yield did not exceed sustainable thresholds, although stocks were declining. Without appropriate management measures, *O. chrysurus* could also become overexploited. Our results emphasize the need to integrate these species into Brazil's fishery management plans to prevent further stock depletion and ensure long-term sustainability.

1 | Introduction

Snapper (family Lutjanidae) are tropical and subtropical marine species found in many parts of the world with high socio-economic commercial and recreational value (Grimes 1987; Duarte and Garcia 1999; Freitas et al. 2011). Historically, snappers were widely exploited for income by many small-scale fishing communities in Brazil, for landings and high market value (Resende et al. 2003). As key predators, snappers are often grouped with similar species like groupers (snapper-grouper complex) due to their life-history traits, trophic ecology, habitat use, and fishery dynamics (Coleman et al. 2000; Heyman 2014). Species within the snapper-grouper complex typically share life-history traits like relatively slow growth, late maturity, and spawning aggregations that contribute to their vulnerability to overexploitation (Chuenpagdee and Pauly 2005; Schärer-Umpierre et al. 2014; França et al. 2021). In Brazil, especially in the Northeast Region, 12 snapper

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1376 **20 YEARS LATER: UPDATED STOCK ASSESSMENT OF SNAPPERS IN**
1377 **NORTHEAST BRAZIL USING AN INTEGRATED STOCK ASSESSMENT**
1378 **FRAMEWORK**

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1380 **Beatrice Ferreira³, George Olavo⁴, Thierry Frédou¹**

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1394 *Abstract*

1395 We updated stock status of *Lutjanus analis*, *Lutjanus jocu*, *Lutjanus synagris*, and
1396 *Ocyurus chrysurus* harvested along the Brazilian northeastern coast. Stock boundaries
1397 were defined according to the Marine Ecoregion classifications at a finer scale, to reflect
1398 population structure of each species. Data were exclusively from the handline fishing
1399 fleet, with removals obtained from fisheries national statistics and published commercial
1400 fishery records. Length compositions were from commercial and scientific surveys.
1401 Abundance indices were estimated from data from three projects and Brazilian official
1402 fishery statistics reports. Growth parameters were from literature and natural mortality
1403 was estimated using the Natural Mortality Tool. Data were analyzed using the Stock
1404 Assessment Continuum Tool. Three of the four species were currently overexploited,
1405 with populations having declined by > 70% from 1950 to 2021. In contrast, *O. chrysurus*
1406 yield did not exceed sustainable thresholds, although stocks were declining. Without
1407 appropriate management measures, *O. chrysurus* could also become overexploited. Our
1408 results emphasize the need to integrate these species into Brazil's fishery management
1409 plans to prevent further stock depletion and ensure long-term sustainability.

1410 **Keywords:** quantitative fisheries, fish management, data-limited, groundfish,
1411 fishery modelling.

1412 *Introduction*

1413 Snapper (family Lutjanidae) are tropical and subtropical marine species found in
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1417 landings and high market value (Resende, Ferreira and Frédou 2003). As key predators,
1418 snappers are often grouped with similar species like groupers (snapper-grouper complex)
1419 due to their life history traits, trophic ecology, habitat use, and fishery dynamics (Coleman
1420 et al. 2000; Heyman 2014). Species within the snapper-grouper complex typically share
1421 life-history traits like relatively slow growth, late maturity, and spawning aggregations
1422 that contribute to their vulnerability to overexploitation (Chuenpagdee and Pauly 2005;

1423 Schärer-Umpierre et al. 2014; França et al. 2021). In Brazil, especially in the Northeast
1424 Region, 12 snapper species are targeted by fishing fleets, with *Lutjanus analis* (mutton
1425 snapper), *L. jocu* (dog snapper), *L. synagris* (lane snapper) and *Ocyurus chrysurus*
1426 (yellowtail snapper) supporting the largest landings (Resende, Ferreira and Frédou 2003).
1427 Adults of these species inhabit rocky bottoms or coral reefs, while juveniles are more
1428 commonly found in coastal waters associated with coastal reefs, seagrass meadows or
1429 estuarine zones and share trophic ecology characteristics, with a diet based mainly on fish
1430 and crustaceans, and to a lesser extent, molluscs and annelids (Lessa, Bezerra Júnior and
1431 Nóbrega 2004; Monteiro, Giarrizzo and Isaac 2009).

1432 In the 1950s, Portuguese fishers introduced vertical longlines, known as
1433 *Pargueiras*, to Brazilian fleets as an alternative to tuna fisheries that were already starting
1434 to decline (Resende, Ferreira and Frédou 2003). Since then, snappers were target species
1435 of several fishing fleets that peaked in the 1980s before declining thereafter (Ivo and
1436 Sousa 1988; Ximenes and Fonteles-Filho 1988). The decline was usually attributed to
1437 depletion of stocks of southern red snapper (*Lutjanus purpureus*), which was the most
1438 fished snapper that was later supplanted by other snapper species. Since the 1990s,
1439 snappers accounted for ~40% of total landings by demersal fisheries in Brazil due to
1440 increased production of other snapper species, including *L. analis*, *L. jocu*, *L. synagris*
1441 and *O. chrysurus*, which were mainly fished in Northeast Brazil (Resende, Ferreira and
1442 Frédou 2003; Olavo 2004; Klippel et al. 2005; Frédou, Ferreira and Letourneur 2009a;
1443 Begossi, Lopes and Silvano 2012). In the early 2000s, the Brazilian government's
1444 REVIZEE project (Lessa, Bezerra Júnior and Nóbrega 2004) conducted stock
1445 assessments of several fish species, including snappers, using age-structured Virtual
1446 Population Analysis (VPA) and the Thompson and Bell Yield-Per-Recruit model (Lessa,
1447 Bezerra Júnior and Nóbrega 2004; Frédou, Ferreira and Letourneur 2009b, 2009a).
1448 Despite identifying overfishing and overexploitation, management plans were never
1449 implemented for these fisheries. To make matters worse, the official Brazilian fisheries
1450 statistical program – ESTATPESCA (IBAMA 2008) ceased to exist in 2007, thereby
1451 stopping the collection and processing of fishery data, especially for coral reef fishes
1452 (Frédou et al 2017; Silva et al 2021; Silva et al 2025).

1453 Some stock assessment approaches use data-limited models that attempt to
1454 maximize use of available limited data (Chrysafi and Kuparinen 2015; Dowling et al.
1455 2019; Cope et al. 2023; Cope 2024a). For example, when data on abundance indices and
1456 age and growth are unavailable or of low quality, data-limited methods can provide ways
1457 to move forward by outlining assumptions and exploring uncertainty (Cope 2024a).
1458 Catch-based models use catch time series that can be combined with species life history
1459 data (e.g., growth equation parameters, maximum length, and age) and assumptions of
1460 stock status to estimate sustainable catches (Cope 2013). Other methods are based on
1461 length frequency distributions (LFD) and life history to estimate sustainable fishing rates
1462 and current relative stock status, such as mortality estimators based on average length that
1463 assume fishing mortality directly influences average length of catches (Gedamke and
1464 Hoenig 2006; Pons, Cope and Kell 2020; Then et al 2018). LFD data is easy and cheap
1465 to collect, so is usually the primary (or only) type of data collected in data-limited
1466 fisheries (Hordyk et al. 2015; Mildemberger, Taylor and Wolff 2017). Because integrated
1467 models can be flexible to available data types, data-limited methods can be built in the
1468 same modelling framework (Cope 2024a). Fisheries management often aims to optimize
1469 yield, while balancing environmental impacts and socioeconomic impacts (Sainsbury,
1470 Punt and Smith 2000; Garcia et al. 2003), so is highly variable and influenced by multiple
1471 factors that can change over time. Additionally, lack of comprehensive data in developing

1472 countries and small-scale fisheries often hinders development of effective management
1473 plans (Cope et al. 2023).

1474 Most fish stocks and associated fisheries in Northeast Brazil, including snappers,
1475 are data-limited, which necessitates alternative assessment methods. We updated stock
1476 assessments for four economically important lutjanid species in Northeast Brazil: *L.*
1477 *analis*, *L. jocu*, *L. synagris*, and *O. chrysurus* using the Stock Assessment Continuum
1478 Tool (SACT) (Cope 2024a, 2024b) an integrative stock assessment model with historical
1479 catch data, length compositions, available abundance indices, and life history. We sought
1480 to determine if stocks of *L. analis*, *L. jocu*, *L. synagris*, and *O. chrysurus* were being
1481 overexploited, and if so, to estimate reference points for fishery harvest management.

1482 *Methods*

1483 *Stock definition*

1484 Stock assessments were defined within boundaries for snapper stocks of the
1485 Brazilian northeastern coast according to defined marine ecoregions (MEs) definitions
1486 (Spalding et al. 2007; Figure 1). This definition was selected based on regional
1487 homogeneity of species composition, habitats, oceanographic characteristics, and fishery
1488 dynamics within the ME of the northeastern Brazilian coast, which has high levels of
1489 biodiversity and is considered a priority for management and conservation (CBD 2014;
1490 Eduardo et al. 2018). The region includes several Marine Protected Areas (e.g. “APA
1491 Costa dos Corais”, “APA Guadalupe”, “APA Barra de Mamanguape”; Ferreira and
1492 Maida 2007). Using MEs as the boundary for stock definition provided an appropriate
1493 proxy for inferring stock boundaries in the Southwest Atlantic and offered more accurate
1494 and finer results than broader systems like Large Marine Ecosystems (LMEs). This was
1495 especially relevant for species like snappers, which may have population structures that
1496 align with finer ecological scales, such as those defined by MEs. Recent research on
1497 otolith shape and chemical composition of *Lutjanus synagris* identified dissimilarities
1498 among populations across different MEs (Dos Santos et al. 2022). However, individuals
1499 did not differ significantly within the Northeast Brazil ME, which reinforced the idea of
1500 a single stock in this area. In contrast, otolith characteristics differed significantly between
1501 Eastern and Northeast Brazil MEs, which suggested these areas should be treated as
1502 distinct management units. Although this evidence suggests separation of stocks between
1503 different MEs, further large-scale studies comparing age composition, growth, mortality,
1504 and reproduction along a latitudinal gradient are required to substantiate a hypothesis of
1505 distinct stocks across these regions.

1506 *Data*

1507 Data were from the commercial handline fishing fleet, which composed more than
1508 95% (IBAMA 2008) of the catches of these species. Fisheries data for snapper
1509 assessments consisted of catch, length-frequency data, and abundance index (Figure 2).
1510 Historical catch data were obtained from official reports on commercial fisheries from
1511 1950 to 2007 (IBAMA 2008), a reconstruction from 2008 to 2015 separated by state
1512 (Freire et al. 2021), and a compilation of catch time series from the Sea Around Us Project
1513 (Sea Around Us 2024), revised and extended by the REPENSAPESCA Project until 2021
1514 (Ferreira, Olavo and França 2022) by geographical units (Northeastern Brazilian states)
1515 and taxonomic units (species) of interest for stock assessment purposes (Figure 3).

1516 Length compositions were from REVIZEE, PRO-ARRIBADA, and
1517 REPENSAPESCA projects. The REVIZEE project during the 1990s collected large
1518 amounts of data from the Brazilian EEZ from industrial fishing fleet based in Natal-BR,
1519 landings from small-scale artisanal fishing fleets from various fishing communities in the

1520 Northeast states, and data from scientific surveys (MMA 2006). Data were collected
1521 monthly from five northeastern Brazilian states: Ceará, Rio Grande do Norte,
1522 Pernambuco, Alagoas, and Bahia (Frédou 2004). The PRO-ARRIBADA project,
1523 implemented in 2008, provided basic information about feeding and reproductive
1524 aggregations of socioeconomically important fish species. The project covered four broad
1525 areas, two of which were within the range of the Northeast Brazil ME and provided length
1526 data from 2009 to 2014. The REPENSAPESCA project, which aimed to assess trends in
1527 the fisheries and the population structure of commercially important reef fish, provided
1528 length composition from 2019 to 2021. Although official commercial fishery records
1529 ceased in 2007, post-2007 sampling continued at the same landing sites using the same
1530 protocols established during the earlier monitoring programs. While this ensured
1531 methodological consistency, representativeness of post-2008 data relative to the broader
1532 commercial fishery cannot be conclusively verified. This limitation was explicitly
1533 recognized in our analysis and interpretation.

1534 Abundance was indexed as catch-per-unit-of-effort (CPUE = kg/day, catch in kg
1535 per day at sea for each vessel). CPUE was estimated using data from the
1536 REPENSAPESCA project (from the state of Rio Grande do Norte), the PRO-
1537 ARRIBADA project (from the state of Pernambuco), and the ESTATPESCA program,
1538 Brazil's official fishery statistics until 2007 (states of Rio Grande do Norte, Paraíba,
1539 Pernambuco, and Bahia). Raw CPUE was used because standardization was not possible
1540 due to lack of factors that can influence catch and effort, such as depth, vessel type, and
1541 spatiotemporal distribution. Diversity of sources and heterogeneity of data, such as
1542 differences in fleet type or fishing area, also hindered standardization CPUE, so
1543 abundance was indexed solely as raw CPUE.

1544 *Life history*

1545 Growth parameters for most species were sourced directly from peer-reviewed
1546 studies within the stock area. However, for *L. analis*, no peer-reviewed literature was
1547 available within the stock area, so growth parameters were estimated using the FishLife
1548 package (Thorson et al. 2017), which provides standardized life-history estimates based
1549 on phylogenetic and ecological covariates. The instantaneous natural mortality coefficient
1550 (M) was estimated for all species using the Natural Mortality Tool (NMT) (Cope and
1551 Hamel 2022) based on the arithmetic mean of four estimators that rely on maximum
1552 observed age: $M = 4.889 * T_{max}^{-0.916}$, $M = \exp(1.717 - 1.01 * \ln(T_{max}))$ (Then et al
1553 2015), $M = 5.4/T_{max}$ (Hamel and Cope 2022), and $M =$
1554 $3k / (\exp(k * ((0.302 * T_{max}) - T_0)) - 1)$ (Alverson and Carney 1975).

1555 *Population dynamics*

1556 Biomass and stock status indicators were estimated using the SACT framework
1557 (Cope 2024b) that uses the Stock Synthesis age-structured modelling framework (Methot
1558 and Wetzel 2013) as a flexible integrated analysis to model variable data availability to
1559 estimate population dynamics (Cope 2013). This approach requires input parameters for
1560 natural mortality (M), von Bertalanffy growth parameters (L_{inf} , K , and t_0), sizes at 50%
1561 and 95% maturity and selectivity, coefficients of the weight-length and fecundity-weight
1562 relationships, parameters of the stock-recruit relationship (steepness and initial
1563 recruitment at the carrying capacity $lnR0$), and catch, length, and (if available) relative
1564 abundance (Figure 4). Due to a lack of data on fecundity at weight, fecundity-weight
1565 coefficients were set to be the same as the weight-length relationship. Steepness (h) was
1566 set to 0.7, the estimated mean value for Perciformes based on the *FishLife* package
1567 (Thorson et al. 2017; Thorson 2020). Initial selectivity inputs were estimated using the

1568 catch-curve method in the *TropFishR* package (Mildenberger, Taylor, and Wolff, 2017),
1569 as preliminary estimates of the size at full selectivity based on observed length-frequency
1570 data. These values were then used to set the initial parameters for the prior distributions
1571 in SACT as part of our modeling approach. The model can estimate selectivity but relies
1572 heavily on these priors in data-limited contexts due to absence of detailed gear-specific
1573 or age-structured data. In our assessment, selectivity was semi-informed and initialized
1574 using empirical estimates that subsequently allowed adjustment within bounds defined
1575 by priors during model fitting. This intermediate approach balanced the need for model
1576 flexibility with limitations of sparse data, to reduce parameter uncertainty while
1577 maintaining biological plausibility. Reference points for spawning stock biomass (*SSB*)
1578 included spawning biomass at maximum sustainable yield (SSB_{MSY}) as a minimum stock
1579 size threshold (MSST) and 125% of SSB_{MSY} as a management target (SSB_{target}). The
1580 instantaneous fishing mortality (F) reference point was the fishing mortality at maximum
1581 sustainable yield (F_{MSY}). Reference points were estimated by SACT based on life-history
1582 parameters provided as model inputs, including growth, natural mortality, maturity, and
1583 length-at-capture, under assumptions of constant selectivity and no explicit stock-
1584 recruitment relationship.

1585 *Assessing uncertainty*

1586 Catch time series were composed of three linked sources: (1) for 1970–2007,
1587 catches were from the official Brazilian program of data collection, ESTATPESCA
1588 (IBAMA, 2008), which compiles all national fishery statistics as the most reliable catch
1589 information; (2) for 1950–1969 and 2008–2015, Freire et al. (2021) reconstructed marine
1590 commercial landings data by systematically addressing gaps and inconsistencies in
1591 Brazil's official fisheries statistics from national statistics, scientific literature, and expert
1592 consultations to estimate unreported catches, discards, and other components absent from
1593 official records; and (3) for 2016–2021, Ferreira et al. (2022) used a similar approach to
1594 extrapolate reconstructions for selected species. Given the lack of official catch statistics
1595 during the last decade, we evaluated sensitivity of stock status estimates to uncertainty in
1596 post-2015 catch data. Specifically, we tested 18 alternative catch scenarios for each
1597 species during 2016–2021, by adjusting removals by $\pm 10\%$ to $\pm 50\%$, in 5%
1598 increments, from a baseline scenario derived by Freire et al. (2021) and Ferreira et al.
1599 (2022), while maintaining the same trend trajectory (Figure 5). These catch trajectories
1600 were used to evaluate robustness of stock status estimates to uncertainty in recent
1601 removals by systematically testing how sensitive model outputs are to reasonable
1602 variations in catch levels. Each scenario was subject to full model fitting and sensitivity
1603 analysis. To integrate results across all scenarios, we applied an ensemble modeling
1604 framework in which each scenario was weighed equally. Stock Synthesis outputs were
1605 processed to extract key management metrics (e.g., fishing mortality, biomass ratios, and
1606 recruitment), constructed probability distributions from their estimates and uncertainties,
1607 and combined them into weighted ensemble trajectories. The equal weighting scheme
1608 reflects our assumption that all scenarios were equally plausible representations of the
1609 system, avoiding subjective prioritization. The ensemble outputs, including mean
1610 trajectories, uncertainty bounds, and Kobe plot quadrant probabilities, provide a more
1611 robust synthesis than individual scenario assessments by accounting for variability across
1612 model structures and assumptions. This approach ensures that management inferences are
1613 derived from a balanced integration of all available evidence. A bootstrap routine was
1614 then used to quantify the probability that each species fell into one of four risk categories,
1615 defined by relative fishing mortality and spawning stock biomass within the Kobe plot
1616 framework. In the green quadrant (lower right), the stock is not overfished and no

1617 overfishing is occurring, which is considered low risk, indicating a sustainable and
1618 healthy stock status. The orange quadrant (upper right) reflects a stock that is not
1619 overfished, but overfishing is occurring, representing moderate risk and signaling
1620 potential need for management action. In the yellow quadrant (lower left), the stock is
1621 overfished, but overfishing is not occurring, also indicating moderate risk. Finally, the
1622 red quadrant (upper left) represents stocks that are both overfished and experiencing
1623 overfishing, a high-risk scenario typically requiring urgent intervention. For the final
1624 model, convergence was tested using the Carvalho et al. (2021) method, which provides
1625 a diagnostic flow chart for model outputs. The first test included checking if the Hessian
1626 matrix was positive and definite, to ensure the function converged to a unique maximum
1627 likelihood estimate. The second test was a jitter analysis of initial values to ensure global
1628 convergence. The third test was for random distribution of residuals that indicated model
1629 assumptions were acceptable. The fourth test was a sensitivity analysis of productivity
1630 parameters to verify consistency of results. The last test was a 5-year retrospective
1631 analysis to verify consistency of results over time.

1632 *Results*

1633 *Diagnostic analysis*

1634 The input parameters used in the Stock Assessment Continuum Tool reference
1635 model for four snapper species (*Lutjanus analis*, *Lutjanus jocu*, *Lutjanus synagris*, and
1636 *Ocyurus chrysurus*) sampled along the northeastern Brazilian coast between 1950 and
1637 2021 are shown in Table 1. These parameters include life history traits such as maximum
1638 age, age and length at maturity, growth parameters from the von Bertalanffy Growth
1639 Function (VBGF), natural mortality rates, weight–length relationships, fecundity
1640 estimates, and the steepness parameter (h) of the stock–recruitment relationship. Data
1641 sources include peer-reviewed studies and the FishLife database, as indicated. Where
1642 confidence intervals (CIs) are available, they are reported alongside the parameter
1643 estimates.

1644 Key stock status indicators for the four snapper species, based on the Stock
1645 Assessment Continuum Tool results, are depicted in Table 2. The unfished spawning
1646 stock biomass (SSB) estimates represent the total reproductive biomass if no fishing had
1647 occurred. The catch at maximum sustainable yield (MSY) is the estimated highest average
1648 annual catch that can be maintained without depleting the stock. The spawning stock
1649 biomass at MSY (SSB_{MSY}) indicates the reproductive biomass level that supports MSY.
1650 Ratios comparing the 2021 spawning stock biomass to SSB_{MSY} (SSB_{2021}/SSB_{MSY})
1651 indicates whether the stock is above or below sustainable levels. Fishing mortality at
1652 MSY (F_{MSY}) is the rate of fishing that achieves MSY, while the ratio of fishing mortality
1653 in 2021 to F_{MSY} (F_{2021}/F_{MSY}) indicates current fishing pressure relative to sustainable
1654 limits. Finally, the stock biomass status categorizes each stock as overexploited or
1655 underexploited and whether overfishing is occurring. These metrics provide a
1656 comprehensive view of stock health and fishing pressure to guide management decisions.

1657 The SACT estimation fitted size composition data well (Figure 6 ; Table 2). For
1658 all four species, diagnostics were within the acceptable range. Final gradients of model
1659 outputs were small (< 0.001), and Hessian matrices for parameters were positive and
1660 definite. Residuals for length compositions and abundance indices were randomly
1661 distributed, except for the length composition of *O. chrysurus* (Figure S1 **Error! Reference source not found.**). All 50 jitter runs for each species were stable despite
1662 variation in initial values, with most runs converging to the same solution as the base
1663 model for each species (Figure S2). For all species, final model outputs for biomass and
1664 fishing mortality were not sensitive to variation in natural mortality or steepness (Figure
1665

1666 S3). Retrospective analyses indicated low bias for most species, except *L. analis*, due to
1667 an imbalance in the sample size among final years (Figure S4).

1668 *Lutjanus analis*

1669 Removals increased until the mid-1970s, and reached a historical peak in 1977,
1670 before declining sharply in the 1980s (Figure 3a). Thereafter, removals increased between
1671 1988 and 1996 and stabilized thereafter. Length composition data for 1996–2000, 2009–
1672 2014 and 2019–2021 were well fit by the model (Figure 6a). The abundance index did
1673 not deviate significantly from the fitted trend but decreased abruptly at the end of the
1674 period (Figure 7a).

1675 Total stock biomass (B) estimates fluctuated from 15,000 t in 1950 to 2,900 t in
1676 2021 (Figure 8a). A steady equilibrium in stock spawning biomass (SSB) between the
1677 early 1980s and 2010 was below B_{MSY} . Relative spawning biomass ranged from 99.6%
1678 at the beginning of the period to 14% in 2021, near B_{target} since the late 1970s, with low
1679 signs of recovery before a sharper decline in the last decade (Figure 8b). Current spawning
1680 biomass was below the limit ($SSB_{2021}/SSB_{MSY} \approx 0.44$). Recruitment fluctuated greatly
1681 in the last two decades (Figure 8c), with peaks in 1999 (>2,200 recruits) and 2014 (~1,900
1682 recruits). Fishing mortality after 1974 surpassed $F/F_{MSY} = 1$, and was lower only briefly
1683 between 1980 and 1994, before rising again until the end of the period ($F_{2021}/F_{MSY} \approx$
1684 2.38; Figure 8d).

1685 *Lutjanus jocu*

1686 Removals increased until the late 1990s, then declined sharply, before plateauing
1687 in 2007 (Figure 3b). Length composition data for 1996–2000, 2009–2014 and 2019–2021
1688 were well fit by the model (Figure 6b). Most length composition data was from the
1689 REVIZEE project in the 1990s, but the last decade was well represented. The abundance
1690 index was only available from 1995 through 2014 and missing for the end of the period
1691 (Figure 7b).

1692 Total biomass was relatively steady, before declining sharply in 1988 and again
1693 in 1996, when it declined ~66%, with a recovery between 1997 and 2002, and a decline
1694 thereafter, when it declined 43% until the end of the period (Figure 9a). Spawning stock
1695 biomass declined to below 30% of initial values, and by the beginning of the last decade,
1696 the stock was below the limit ($SSB_{2021}/SSB_{MSY} \approx 0.82$; Figure 9b). Recruitment was
1697 dynamic throughout the period, with large recruitments in 1994, 1995, and 1996 (Figure
1698 9c). Fishing mortality increased steadily until the last decade, to surpass MSY (Figure
1699 9d), but was at sustainable levels by the end of the period ($F_{2021}/F_{MSY} \approx 1$).

1700 *Lutjanus synagris*

1701 Removals were below 500 tons in early decades, before rising above 1500 tons
1702 between 1992 and 2008, where it stayed thereafter (Figure 3c). Length compositions were
1703 based on low sample sizes (Figure 6c) and the abundance index was only available before
1704 2010 (Figure 7c).

1705 Total biomass declined steadily since the beginning of the period, increasingly
1706 after the 1970s (Figure 10a), with a partial recovery between 1998 and 2003, and a decline
1707 thereafter. Spawning stock biomass first dropped below MSY at the end of the 1990s,
1708 recovered until 2015, and declined sharply to low levels thereafter (Figure 10b)
1709 ($SSB_{2021}/SSB_{MSY} \approx 0.78$). Recruitment was relatively steady, with peaks in 1997, 2007,
1710 and 2016 (Figure 10c). Fishing mortality was constant until 1990, but increased regularly
1711 thereafter, to surpass F_{MSY} after 2000, and was dangerously high by the end of the period
1712 (Figure 10d) ($F_{2021}/F_{MSY} \approx 2.59$).

1713 *Ocyurus chrysurus*

1714 Removals increased since the start of the period, especially after 1972, and
1715 surpassed 2000 tons in 2003 and remained stable since then (Figure 3d). Length
1716 compositions were from 1996 to 2000 and 2009 to 2014, and lacking thereafter, except
1717 in 2020 (Figure 6d). The abundance index was available from 1995 to 2011 and missing
1718 thereafter (Figure 7d).

1719 Total biomass decreased steadily between the start of the period and 1996, with a
1720 peak between 1997 and 2001, after which the stock sharply declined until 2014, when it
1721 stabilised until the end of the period (Figure 11a). Spawning stock biomass was never
1722 below MSY but crossed the B_{target} into the precautionary zone ($SSB_{2021}/SSB_{MSY} \approx$
1723 1.42) in the last decade (Figure 11b). Recruitment increased steadily, with peaks in 1993,
1724 1998, and 2004 (Figure 11c). Fishing mortality was below the reference point of
1725 $F/F_{MSY} = 1$ in the entire period but increased in the last decade ($F_{2021}/F_{MSY} \approx 0.86$;
1726 Figure 11d).

1727 *Uncertainty*

1728 For each species, different landings scenarios converged to the same final status.
1729 *L. analis* had an 88.9% probability of being overfished and experiencing overfishing, and
1730 an 11.1% probability of being overfished but not experiencing overfishing (Figure 12a).
1731 *L. synagris* had a 96.6% probability of being overfished and experiencing overfishing,
1732 and a 3.1% probability of experiencing overfishing but not overfished (Figure 12b). *L.*
1733 *jocu* had a 49.6% probability of being overfished and experiencing overfishing, a 43.3%
1734 probability of being overfished but not subject to overfishing, and a 7.1% probability of
1735 being neither overfished nor undergoing overfishing (Figure 12c). *O. chrysurus* had a
1736 57.0% probability of being neither overfished nor subject to overfishing, and a 39.0%
1737 probability of experiencing overfishing but not overfished (Figure 12d). Stock status
1738 estimates were generally robust to uncertainty in catch data. Consistency among
1739 scenarios, particularly for *L. analis* and *L. synagris*, indicated stability in assessment
1740 outcomes despite uncertainty in catch assumptions.

1741 *Discussion*

1742 Ours is the first formal and comprehensive stock assessment of all four snapper
1743 species on the northeast coast of Brazil in the past 20 years. Our primary objective was
1744 to highlight the critical need for science-based management of demersal fisheries in the
1745 region. Our results revealed clear evidence of prolonged overexploitation of some of the
1746 assessed stocks, which has yet to be addressed by any formal management framework in
1747 the region but highlights an urgent need for a regulatory framework to prevent further
1748 depletion of these stocks. By updating the status of these key fisheries, we aimed to
1749 facilitate the integration of scientific assessment into regional management planning to
1750 stimulate adoption of sustainable harvest strategies. Our study, therefore, is pivotal for
1751 bridging the gap between scientific assessment and development of active management
1752 policies in the region. Although assessments of other fisheries resources have been
1753 completed, such as shrimp (Barros et al. 2021; Peixoto et al. 2021; Aragão et al. 2022),
1754 recent assessments of demersal fish in the North and Northeast regions are rare (Ferreira,
1755 Olavo and França 2022). While the REPENSAPESCA project preliminarily assessed the
1756 status of these species, it relied on the Large Marine Ecosystem (LME16) approach,
1757 which may have led to mixed signals due to the potential presence of distinct stocks in
1758 different regions. In contrast, we used marine ecoregions (MEs), which are considered a
1759 more appropriate boundary for stock definition, for finer-scale and more accurate
1760 assessments. Using historical catch and length data and abundance indices, we evaluated

1761 the status of the four main socioeconomically important snapper species caught along the
1762 northeastern coast of Brazil. Our results indicated that three of the four stocks have been
1763 overexploited for at least a decade, with estimated fishing mortality still much higher than
1764 sustainable levels. This suggests that these stocks are currently being overexploited.

1765 An advantage of Stock Assessment Continuum Tool is flexibility in handling
1766 multiple types of data, including catch data, length compositions, abundance indices and
1767 life-history (LH) information, which is particularly valuable for data-limited fisheries,
1768 such as ours (Methot and Wetzel 2013), to allow for a more refined and data-informed
1769 assessment of stock status (Rudd et al. 2021). This flexibility allows the evaluation of
1770 how different data types contribute to model outputs and where key assumptions, such as
1771 selectivity or mortality, drive uncertainty (Cope et al 2023; Cope 2024a; Rudd et al 2021).
1772 This allows practitioners to identify which inputs are most informative and where
1773 improved data collection would reduce model uncertainty. However, depending on inputs
1774 provided and assumptions made, the model output was still data-limited, even when using
1775 a complex modelling framework (Cope 2024a). One particular limitation of SACT for
1776 data-limited fisheries was selectivity (Cope 2024a; Cope 2024b). For data-limited
1777 fisheries like ours, the lack of data on gear- and size-specific selectivity necessitates the
1778 use of simplified or assumed selectivity curves, which may introduce bias in model
1779 outputs if the true selectivity differs significantly from these assumptions (Hovgard and
1780 Lassen 2000; Cope 2024a). As a result, model estimates of fishery mortality, biomass,
1781 and stock status be sensitive to assumptions, thereby underscoring the need for caution
1782 when interpreting results and for future efforts to improve empirical data on gear
1783 performance and size-at-capture (Chen, Chen and Stergiou 2003; Hoshino et al 2014).
1784 Thus, while SACT offers a comprehensive approach, the quality of output is influenced
1785 by quality and breadth of available data (Cope 2024a). Our study also demonstrated the
1786 importance of incorporating sensitivity analyses to more accurately characterize
1787 uncertainty in model specifications (Tagliarolo, Cope and Blanchard 2021).

1788 Results of our assessment using the Stock Assessment Continuum Tool confirmed
1789 stock assessments of *L. analis*, *L. jocu*, and *L. synagris* as overfished and experiencing
1790 overfishing more than 20 years ago (Frédou, Ferreira and Letourneur 2009a, 2009b).
1791 Previous assessments used traditional age-structured methods, Virtual Population
1792 Analysis (VPA) and Yield-Per-Recruit models, indicated that *L. analis*, *L. jocu*, and *L.*
1793 *synagris* were overfished and subject to overfishing, whereas *O. chrysurus* was under
1794 significant exploitation pressure (Frédou et al. 2009a). Consistency among modeling
1795 approaches and assessment periods suggests a persistent pattern of overexploitation for
1796 these species that reinforces the need for management intervention. In contrast, our
1797 assessment of *O. chrysurus* stock status (not overfished and not undergoing overfishing)
1798 differed from the earlier assessment (overexploited), perhaps because more recent data
1799 used in our assessment, despite gaps and limitations, captured changes in fishing effort or
1800 stock recovery since the earlier assessment. Alternatively, methodological differences
1801 between the earlier assessment (Frédou et al. 2009a) and ours may have influenced how
1802 exploitation status was interpreted, particularly in the absence of fishery-independent
1803 data. Finally, shifts in fleet behavior, species targeting, or market dynamics may have
1804 altered fishing pressure on *O. chrysurus* relative to other species. This historical
1805 comparison supports the credibility of the current assessment, while emphasizing the
1806 chronic nature of overfishing for most snapper stocks and also highlights the value of
1807 regular stock assessments using updated methods and data to track changes in exploitation
1808 status over time, to guide the development of effective, adaptive management strategies.

1809 Table 2 highlights important contrasts in the exploitation status and fishing
1810 pressure across the four snapper species studied along the northeastern Brazilian coast.

1811 Notably, *Lutjanus analis* and *Lutjanus synagris* exhibited clear signs of overexploitation,
1812 characterized by spawning biomass levels below sustainable targets and fishing mortality
1813 rates substantially exceeding biological reference points. For *L. synagris*, these results
1814 closely align with findings from García-Caudillo et al. (2024) for the southern Gulf of
1815 Mexico, where the species was also found to be overfished, highlighting similar pressures
1816 across different parts of the species' range and reinforcing concerns about stocks'
1817 depletion. This combination indicates that these stocks are under significant stress from
1818 current fishing activities, and without effective management interventions, their recovery
1819 may be compromised. The situation calls for management measures such as closed
1820 seasons or closed fishing areas to alleviate fishing pressure and allow rebuilding of
1821 reproductive capacity (Shertzer et al. 2024). *Lutjanus jocu*, meanwhile, presents a
1822 relatively more moderate picture. Although its spawning biomass is somewhat depleted,
1823 fishing mortality is only slightly above sustainable levels. This suggests that while the
1824 stock is vulnerable, it may respond positively to management aimed at reducing fishing
1825 effort to avoid further depletion, as seen in other cases (SEDAR 2021; Bachelet et al.
1826 2025), and continued monitoring will be critical to detect trends and ensure that
1827 exploitation remains within safe biological limits. In contrast, *Ocyurus chrysurus* stands
1828 out as currently underexploited, with spawning biomass above sustainable reference
1829 points and fishing mortality below the threshold associated with maximum sustainable
1830 yield. These patterns mirror those observed in García-Caudillo et al. (2024), where the
1831 species was also found to be in relatively healthy condition under a precautionary
1832 assessment framework. This relatively favorable status offers an opportunity to
1833 implement precautionary management practices that maintain stock productivity and
1834 prevent overfishing before it becomes a concern (FAO 1995; Fisheries and Oceans
1835 Canada 2009). In conclusion, managing snapper fisheries in the Northeast of Brazil
1836 requires a multifaceted approach that integrates ecosystem considerations, climate change
1837 adaptation, precautionary principles, and strong stakeholder involvement. The present
1838 assessment underscores the critical role of improved data collection to enhance the
1839 accuracy and reliability of stock assessments to support evidence-based management. Our
1840 assessment highlights the limitations imposed by inconsistent or missing data,
1841 particularly in later years of the assessed period, which limits detecting recent changes in
1842 fishing pressure, recruitment, and biomass. Establishing a systematic and continuous data
1843 collection program, including standardized catch records, biological sampling of length,
1844 age, and maturity, and fishery-independent surveys, is critical for reducing uncertainty in
1845 model estimates and strengthening the scientific basis of future assessments and the
1846 development of more responsive and precautionary management strategies (FAO 2005).
1847 Lack of a management response is a significant challenge in fisheries management in
1848 Northeast Brazil. Effective fisheries management requires integration of peer-reviewed
1849 science into decision-making processes, moving beyond academic stock assessments
1850 toward actionable management advice (Hilborn 2011). This highlights a need for more
1851 responsive and dynamic management that uses scientific data to foster collaboration
1852 among researchers, policymakers, and stakeholders. Strengthening the link between stock
1853 assessments and management actions through comprehensive and adaptive strategies will
1854 be crucial for ensuring the long-term sustainability of fish stocks and the communities
1855 that depend on them (Alcock 2004). Based on model results for *L. analis*, *L. jocu*, and *L.*
1856 *synagris*, these three fisheries are at risk if current fishing effort continues. This serves as
1857 an example of small-scale fisheries struggling due to neglect, with no consistent data
1858 collection or stock assessments. For *O. chrysurus*, although the fishery is currently at safe
1859 levels, increasing fishing pressure could lead to a precarious situation without proper
1860 management measures. Therefore, catch accounting, data collection, and accurate life

1861 history estimates are needed to reduce uncertainty in results of stock assessments for these
1862 species, so we recommend these four species be prioritized for inclusion in future
1863 Brazilian fisheries management plans, including continuous on-site monitoring (i.e.,
1864 recording landings and discards), control, and surveillance of the fisheries, along with
1865 continuous collection of biological samples, as well as for other species that share similar
1866 life history traits or are associated with similar demersal fisheries on the Northeast coast.
1867 The protocol followed in this work, including life-history parameters estimation, stock
1868 assessment, and sensitivity analysis, can be replicated for other species of the Brazilian
1869 Northeast coast and elsewhere.

1870 *Broader Management Implications*

1871 Management of snapper stocks in Northeast Brazil cannot be considered in
1872 isolation because fisheries are multispecies and of low selectivity (Lessa 2006; Lucena-
1873 Frédou et al. 2021). Gear used by these fisheries often captures multiple species, so
1874 management must be part of an interconnected system, rather than focusing on single-
1875 species stock assessments (Frédou 2009b). This multi-species perspective aligns with the
1876 principles of Ecosystem-Based Fishery Management (EBFM), which advocates for a
1877 holistic approach that incorporates ecological relationships, such as predator-prey
1878 dynamics, habitat dependencies, and species interactions, into management plans (Garcia
1879 2003). The four assessments presented here enhance understanding of the status of
1880 snapper stocks in Northeast Brazil, where stock assessments and data-driven management
1881 are limited. Our results are intended to inform ongoing discussions within local fisheries
1882 co-management councils and governmental bodies, such as the Brazilian Ministry of
1883 Fisheries (MPA) and the Brazilian Institute of Environment and Renewable Natural
1884 Resources (IBAMA), particularly as Brazil advances toward more structured and
1885 adaptive fishery management. Given the multispecies nature of these fisheries and data
1886 gaps in the region (MMA 2006; Lessa, Bezerra Júnior and Nóbrega 2004), this
1887 information can aid in prioritizing management actions, particularly when addressing
1888 data-poor fisheries. Currently, no harvest control rules, or species-specific management
1889 plans exist for these stocks, so stock status information is essential for prioritizing
1890 management actions (Frédou et al 2017; Silva et al 2021; Silva et al 2025). In the absence
1891 of data-rich assessments, our results can support the implementation of precautionary
1892 measures, including spatial or seasonal closures and participatory monitoring strategies.
1893 In addition to ecosystem complexities, climate change must be integrated into
1894 management strategies to ensure long-term sustainability of snapper fisheries. Effects of
1895 climate change, such as ocean warming, changes in current patterns, and acidification are
1896 likely to alter fish distributions, reproductive cycles, and habitat suitability (Cheung et al.
1897 2009; Petitgas et al. 2013; Sydeman et al. 2015). Such changes could exacerbate
1898 challenges already faced by multispecies fisheries, highlighting the need to incorporate
1899 climate-related variables into stock assessments and management plans (Vinther et al
1900 2004; Dolan et al 2016). For instance, future management strategies could include
1901 adaptive measures that allow for flexible response to shifting species distributions or
1902 altered ecosystem productivity, to ensure management is still relevant under changing
1903 environmental conditions.

1904 Our findings are relevant for future applications, such as Marine Spatial Planning
1905 (MSP) or Management Strategy Evaluation (MSE) frameworks tailored for the region.
1906 Adopting a precautionary approach within an adaptive management framework is crucial
1907 for addressing ecological and environmental uncertainties by promoting conservative
1908 catch limits and allowing for adjustments as new data become available to ensure
1909 management remains responsive to real-time ecosystem changes (Rodriguez-Perez et al

1910 2023). MSE can be valuable for enabling testing of management strategies and, by
1911 simulating possible outcomes, it can identify the most effective strategies under
1912 conditions of uncertainty, as a key component of adaptive fisheries management in
1913 dynamic, data-poor fisheries (Butterworth 2007, 2008), like those in the Northeast of
1914 Brazil. However, effective fisheries management is not solely about ecological and
1915 environmental considerations but must also account for socioeconomic realities of
1916 communities that depend on these resources (Garcia 2003; Rodriguez-Perez 2023).
1917 Snapper fisheries are vital to the livelihoods of many local fishers (Ivo and Sousa 1988;
1918 Resende, Ferreira and Frédoú 2003; Frédoú, Ferreira and Letourneur 2009a), so
1919 management strategies need to balance ecological sustainability with socioeconomic
1920 well-being. Co-management practices that involve collaboration among government
1921 agencies, fishers, and local communities offer a pathway to achieving this balance. By
1922 involving stakeholders directly in decision-making, co-management can foster local
1923 stewardship, improve compliance with regulations, and ensure management is grounded
1924 in local knowledge and experience, while support for alternative livelihoods during closed
1925 seasons or stock recovery periods can help alleviate economic pressures on fishers to
1926 ensure short-term sacrifices lead to long-term benefits (Berkes 2009; Motta et al 2022).

1927 To integrate these considerations effectively, MSP and other management options
1928 like the Precautionary Approach (Dowling et al. 2019; Ono, Langangen and Stenseth
1929 2019; Mildenerger et al. 2022) are valuable tools that can be employed. MSP facilitates
1930 the organization of ocean space to minimize conflicts and promote sustainable resource
1931 use by identifying and protecting critical habitats. A spatially explicit approach is
1932 particularly relevant for multispecies fisheries, such as snapper fisheries, by helping to
1933 safeguard essential fish habitats and optimize the allocation of ocean resources (Frédoú
1934 2009b). By combining MSP with precautionary and adaptive management strategies,
1935 fisheries management in the Northeast of Brazil can become more resilient to
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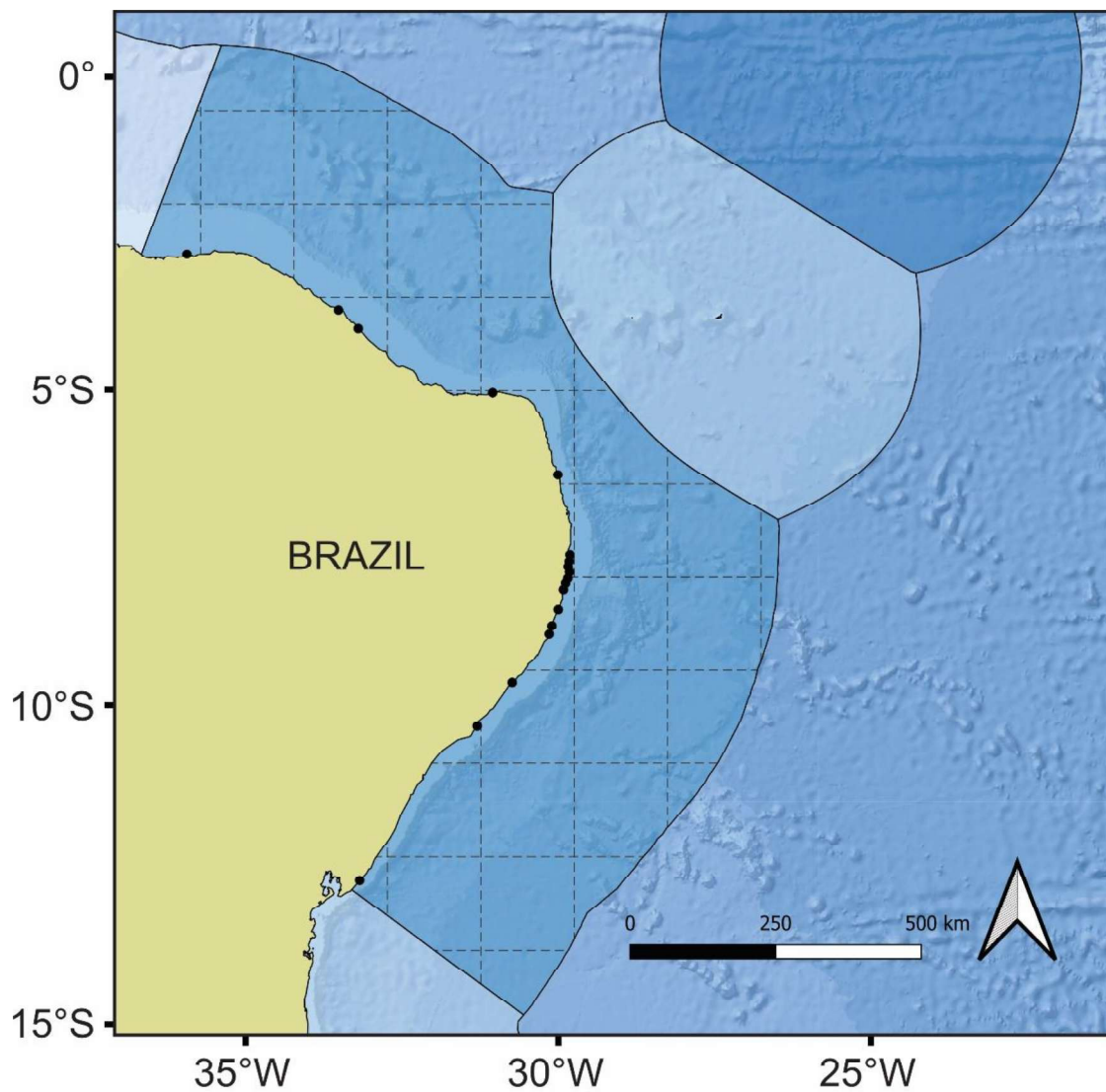
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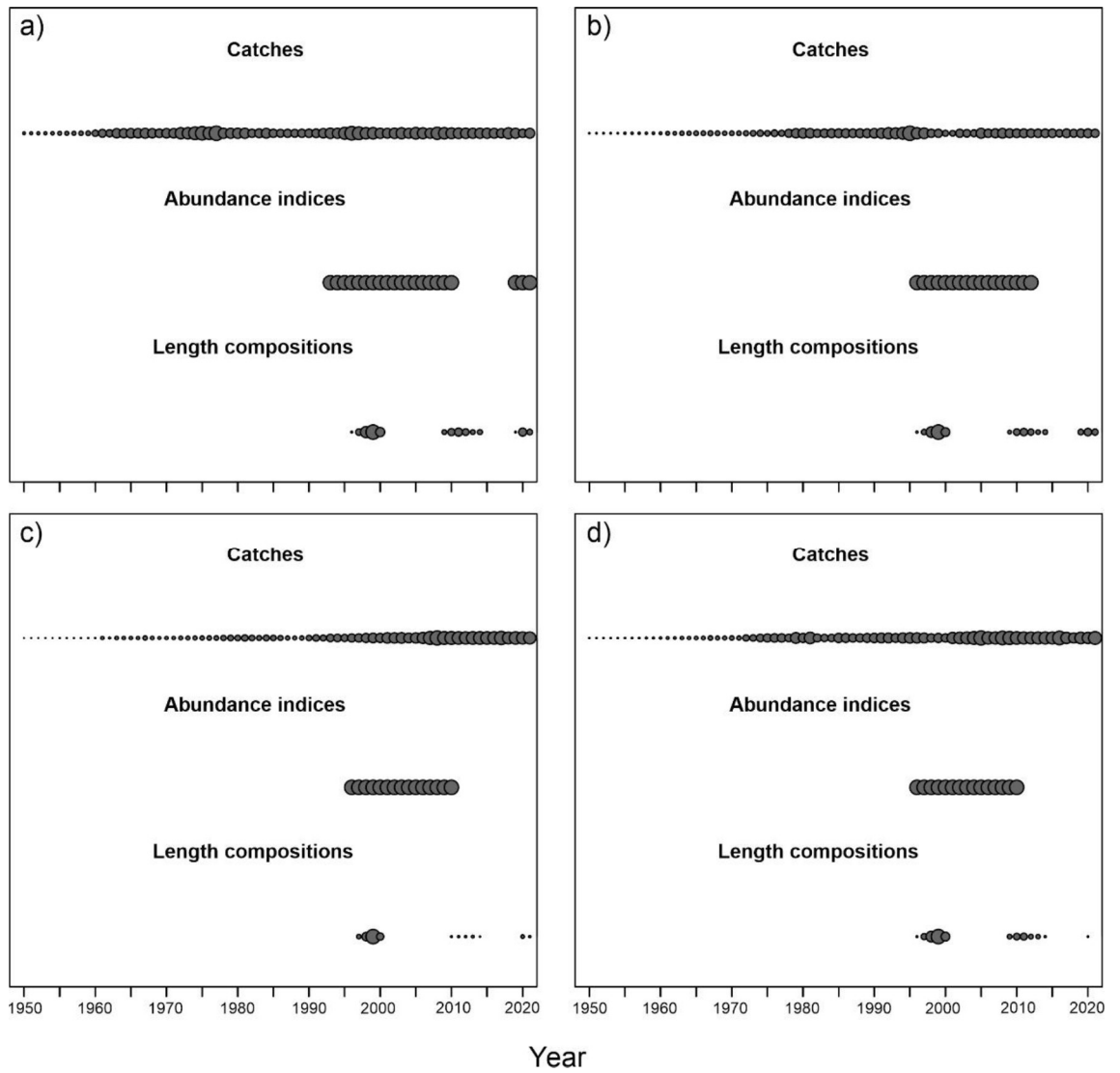
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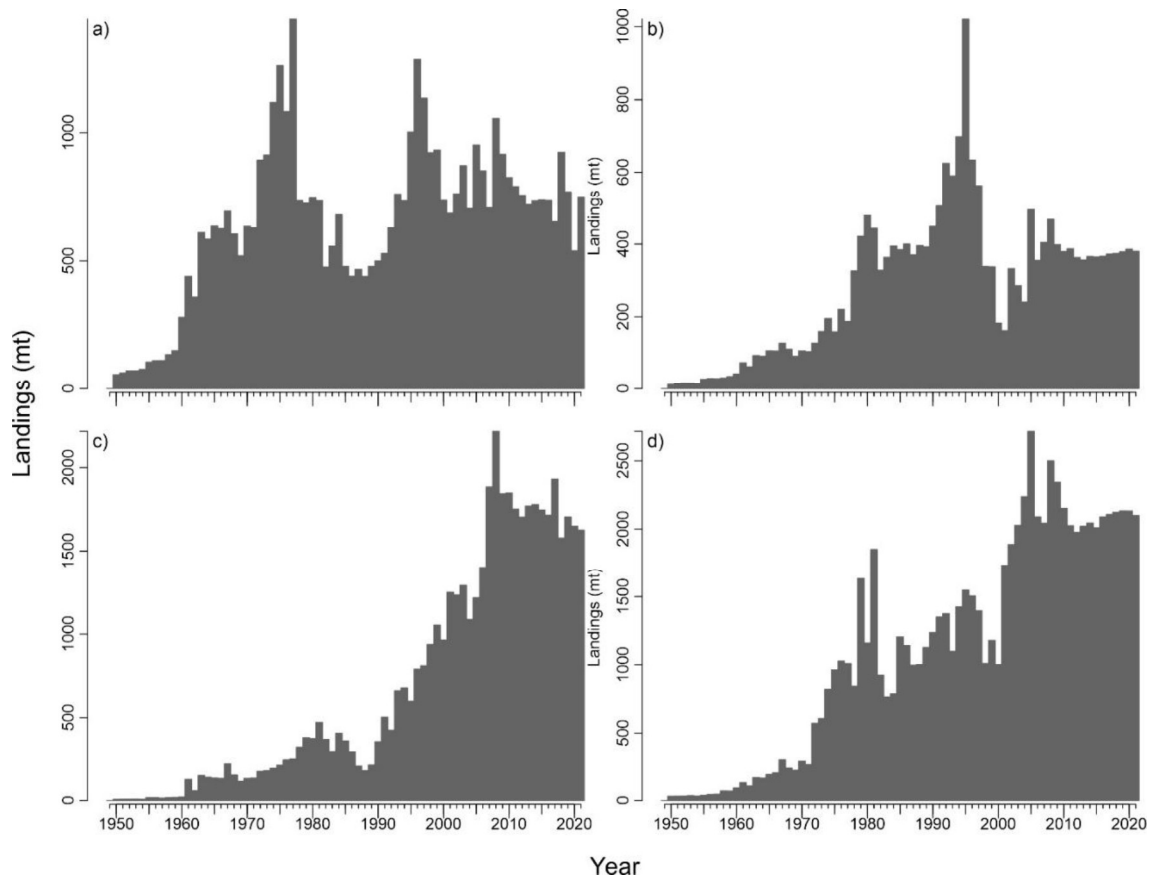
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2204 Figure 1: Study area with delimitations of Marine Ecoregions on the Northeastern coast of
2205 Brazil. The cross-hatched area represents the snapper stock area in the Northeast Brazil ME.
2206 Black dots along the coast represent landing ports.



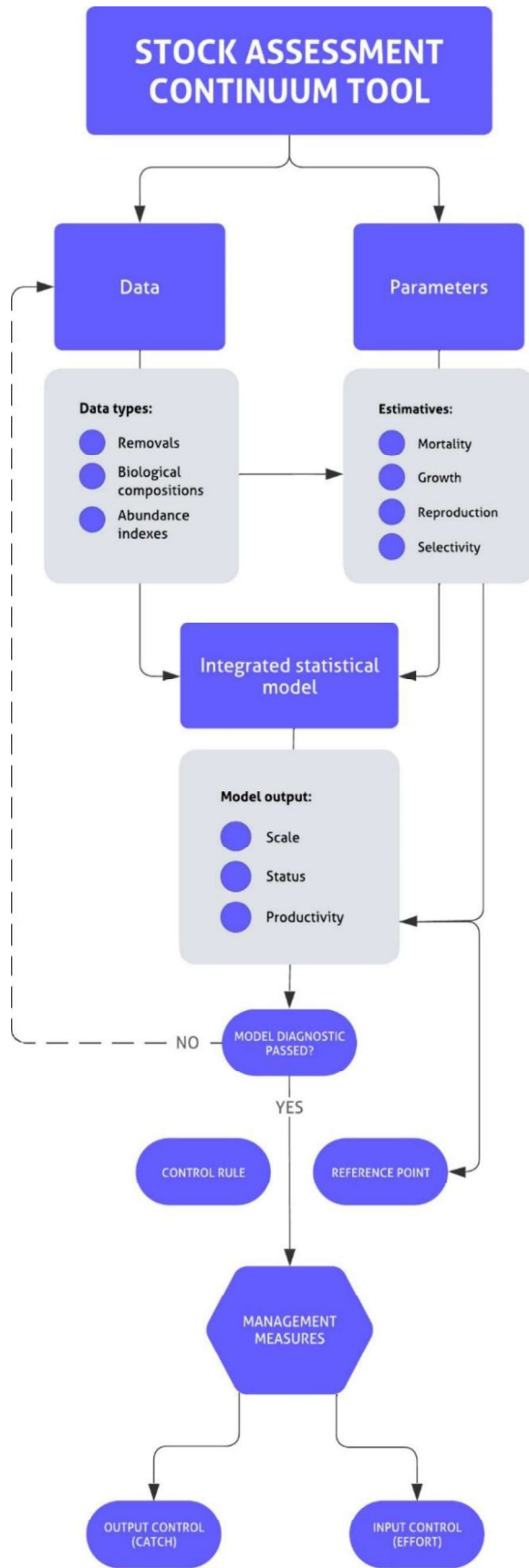
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2208 Figure 2: Catches, abundance indices, and length compositions used in stock assessments of (a)
 2209 *Lutjanus analis*, (b) *Lutjanus jocu*, (c) *Lutjanus synagris*, and (d) *Ocyurus chrysurus* along the
 2210 northeastern Brazilian coast during 1950–2021. Bubble sizes are scaled to represent the
 2211 absolute values of catches, abundance indices, and the number of individuals measured for
 2212 length compositions.



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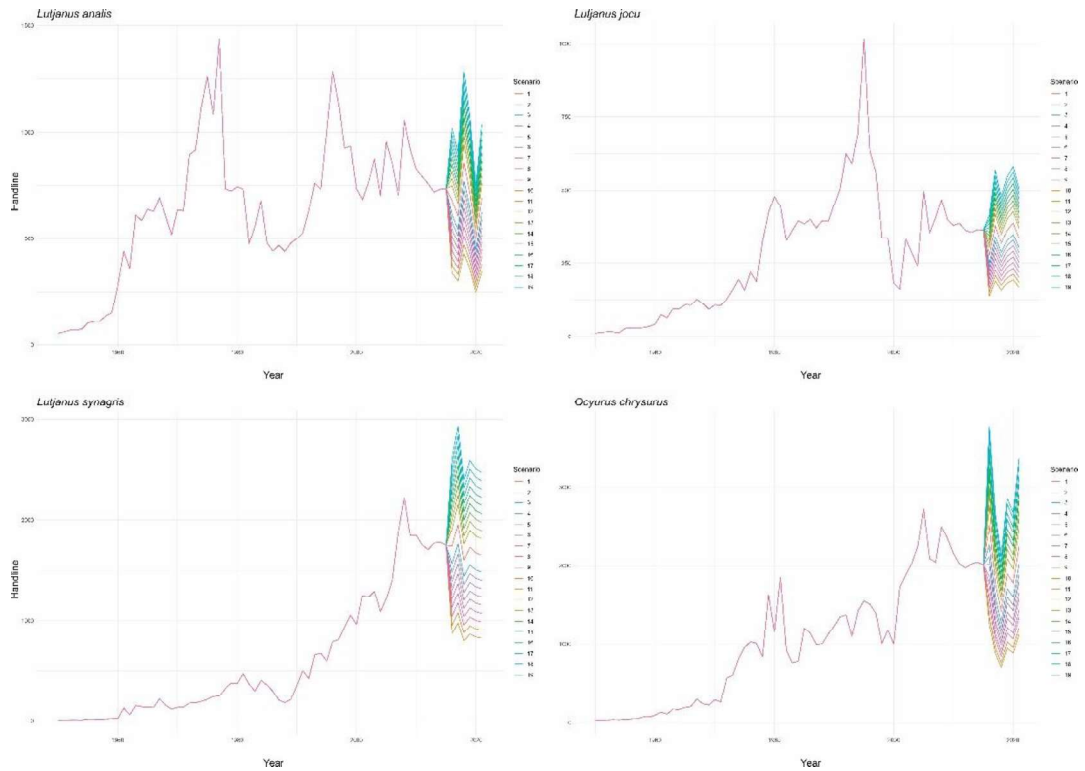
2214 Figure 3: Landings of (a) *Lutjanus analis*, (b) *Lutjanus jocu*, (c) *Lutjanus synagris*, and (d)
 2215 *Ocyurus chrysurus* along the northeastern Brazilian coast during 1950–2021.



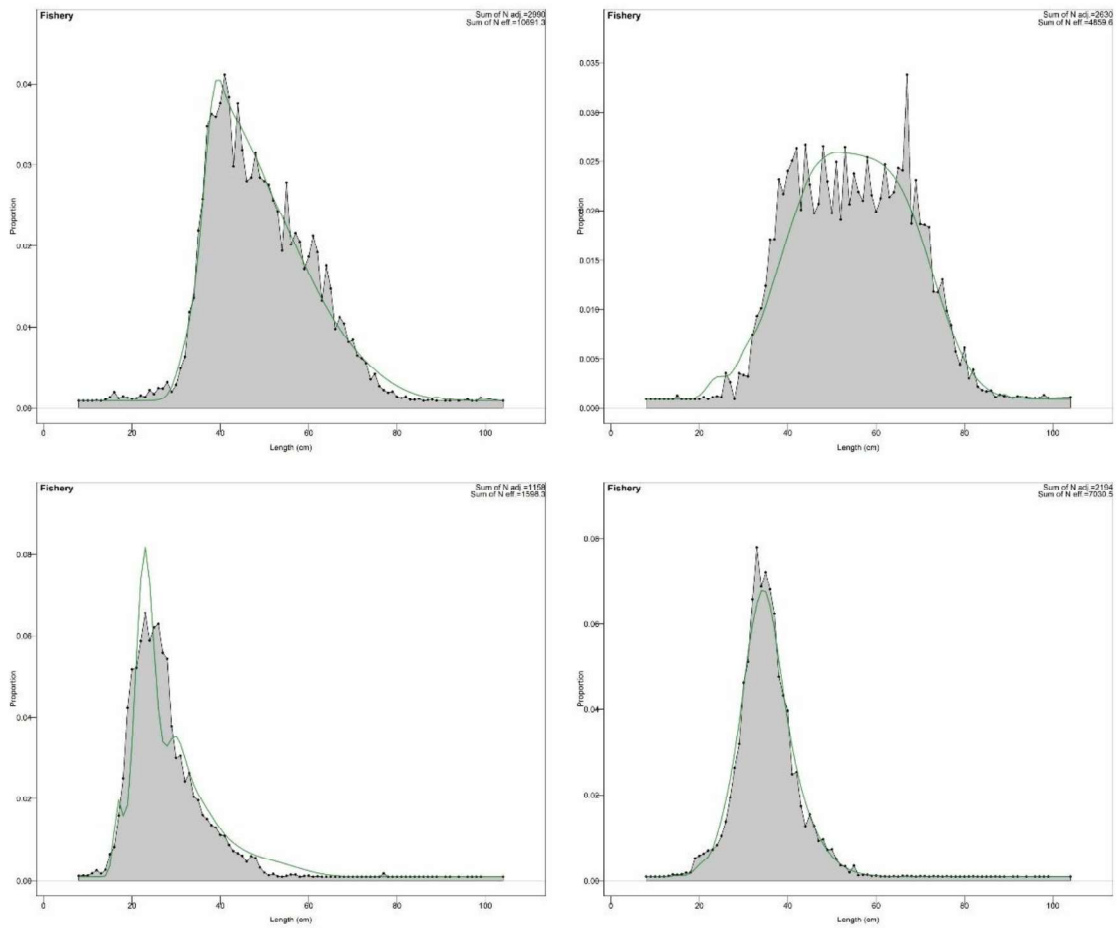
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2217 Figure 4: Flow chart illustrating the Stock Assessment Continuum Tool work path used to
 2218 assess status of stocks of (a) *Lutjanus analis*, (b) *Lutjanus jocu*, (c) *Lutjanus synagris*, and (d)
 2219 *Ocyurus chrysurus* along the northeastern Brazilian coast during 1950–2021. solid arrows

2220 indicate guidance regarding flow of the process from the top to the bottom. The dotted line
 2221 indicates a detour path if diagnostic tests fail.

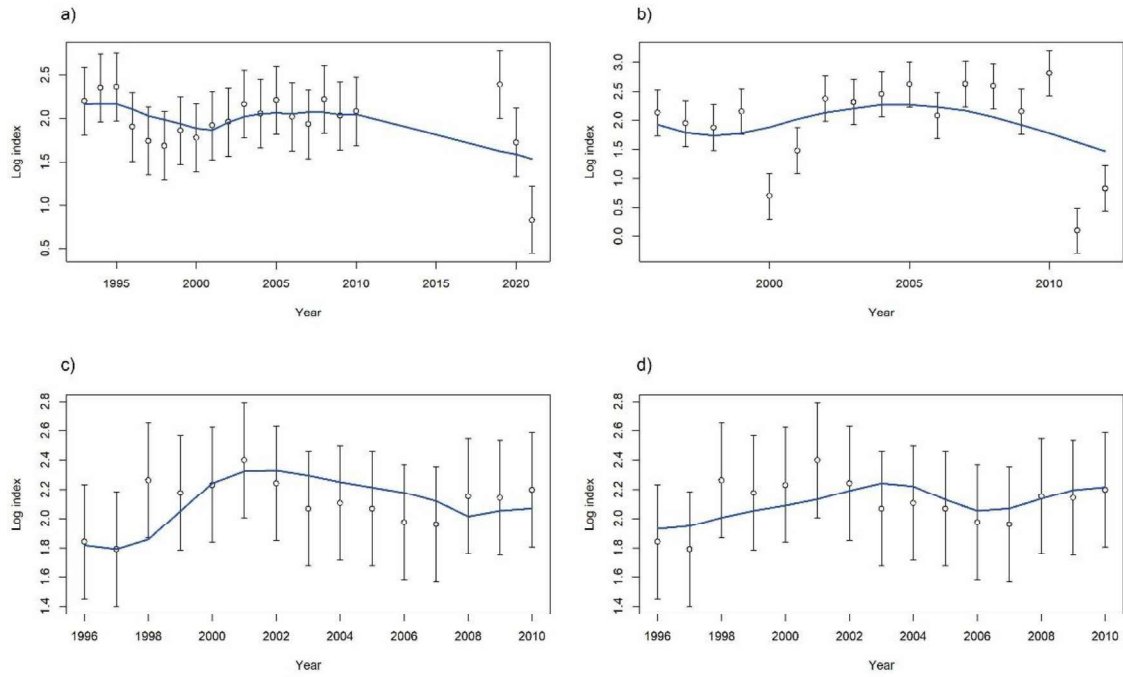


2222
 2223 Figure 5: Alternative catch scenarios for 2016–2021 used in sensitivity analyses of stock status
 2224 estimates of (a) *Lutjanus analis*, (b) *Lutjanus jocu*, (c) *Lutjanus synagris*, and (d) *Ocyurus*
 2225 *chrysurus* along the northeastern Brazilian coast during 1950–2021. Each line represents a
 2226 hypothetical catch trajectory, scaled by $\pm 10\%$ to $\pm 50\%$ (in 5% increments) relative to the base
 2227 scenario while preserving the original scale.



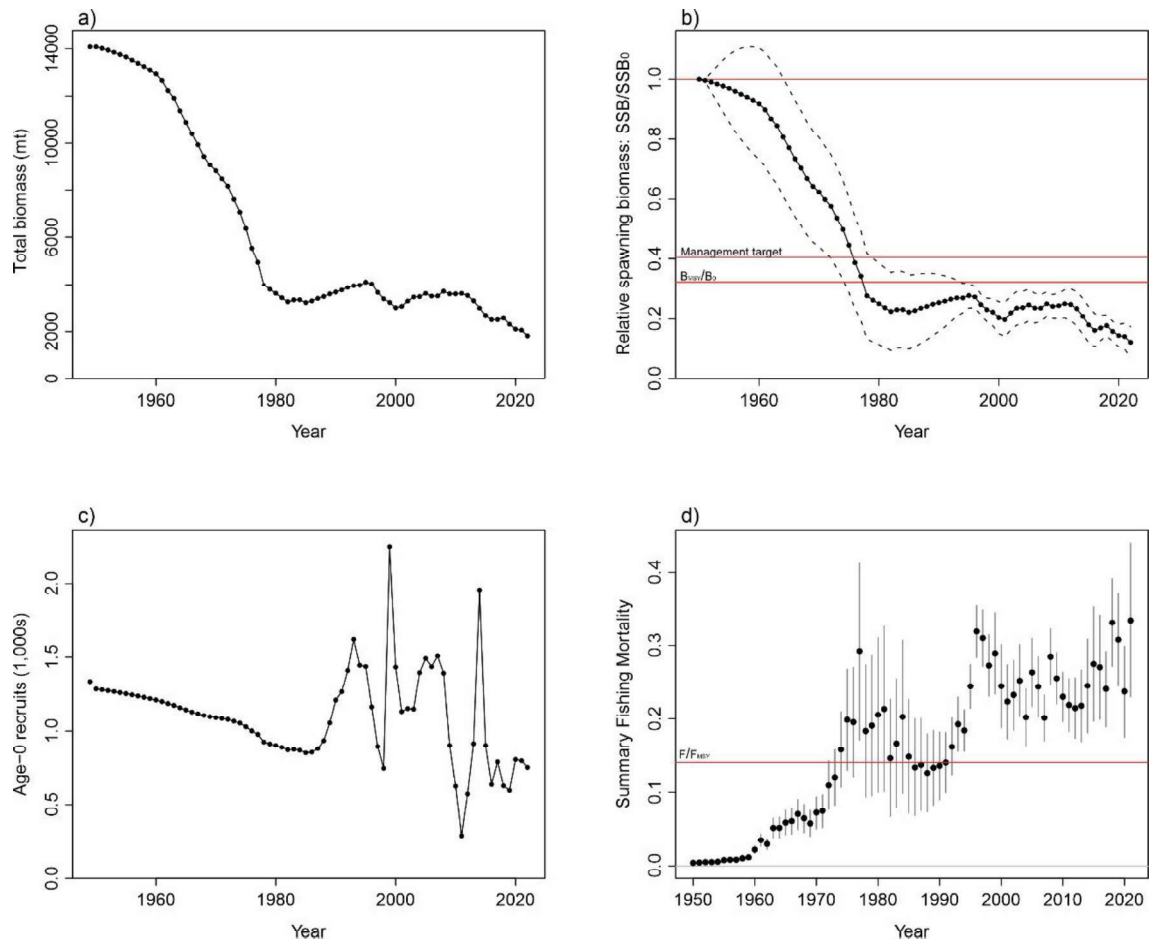
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2229 Figure 6: Length distributions of (a) *Lutjanus analis*, b) *Lutjanus jocu*, c) *Lutjanus synagris*, and
 2230 d) *Ocyurus chrysurus* sampled along the northeastern Brazilian coast during 1950–2021. Dots
 2231 indicate observed values, and green lines indicate model estimates.



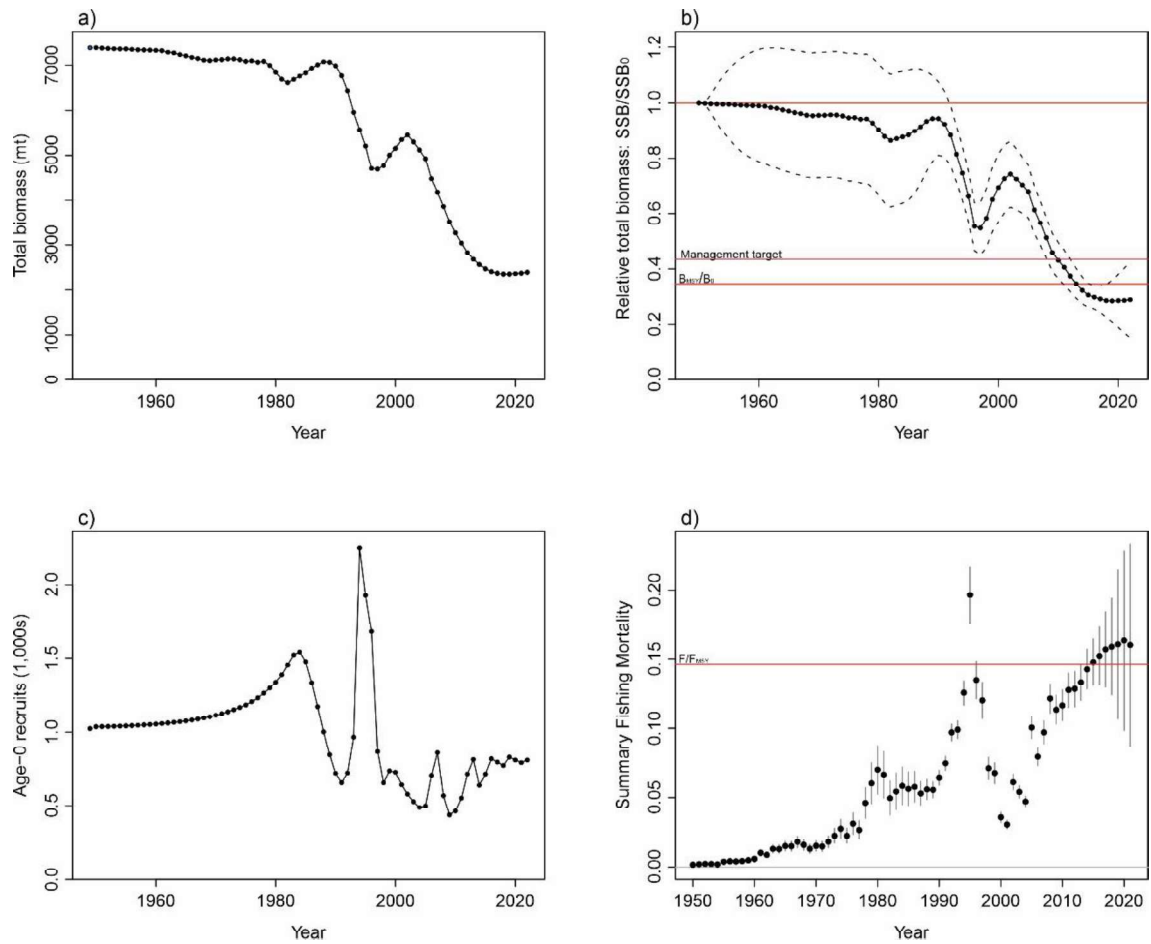
2232

2233 Figure 77: Index of abundance of a) *Lutjanus analis*, b) *Lutjanus jocu*, c) *Lutjanus synagris*, and
 2234 d) *Ocyurus chrysurus* sampled along the northeastern Brazilian coast during 1950–2021. Dots
 2235 represent observed values, blue lines indicate model estimates, and vertical lines indicate
 2236 coefficient of variation.



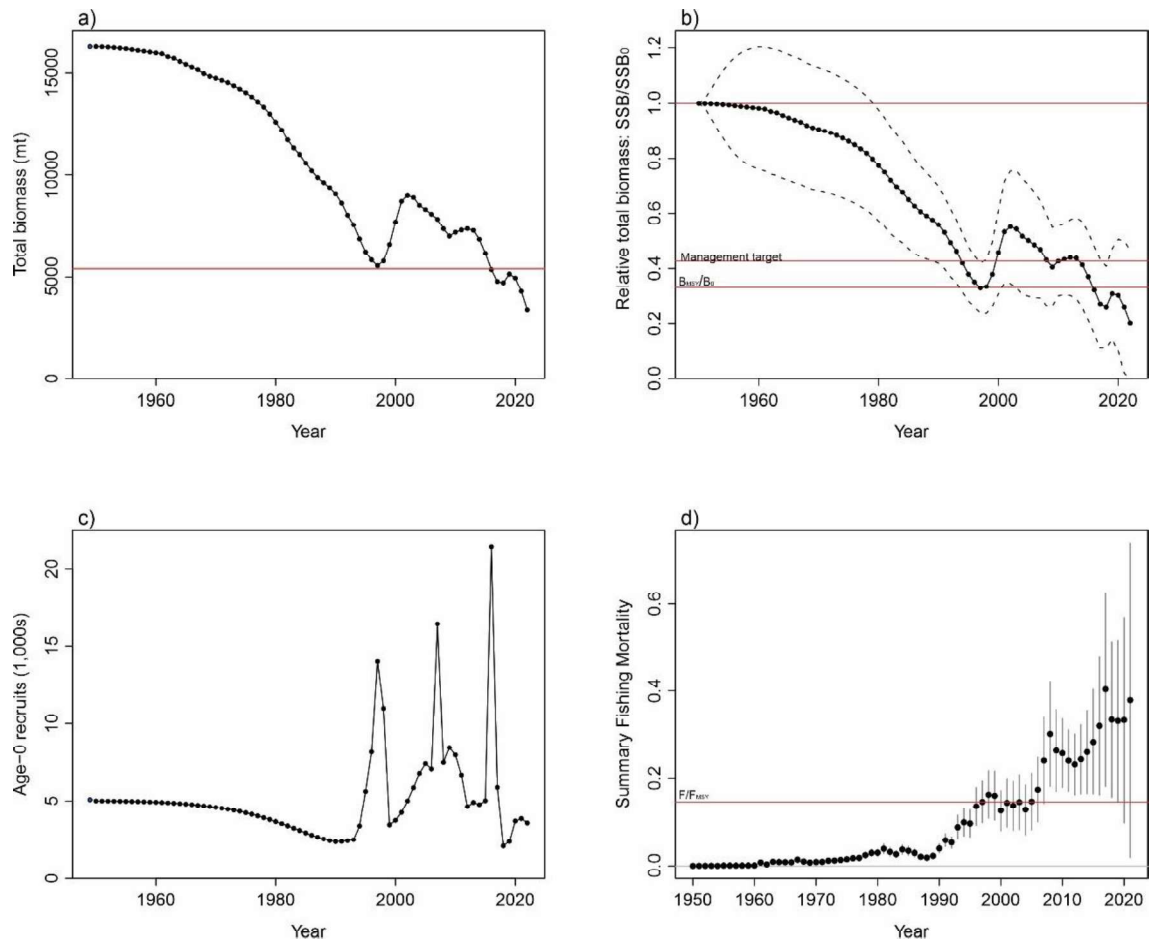
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2238 Figure 88: Spawning stock biomass (SSB) (a), relative spawning biomass SSB/SSB_0 (red
 2239 horizontal lines represent relative SSB at MSY and the management goal of 25% above MSY)
 2240 (b), age-0 recruitment (c), and fishing mortality (red horizontal line represents fishing mortality
 2241 at MSY) (d) estimated by the Stock Assessment Continuum Tool for *Lutjanus analis* along the
 2242 northeastern Brazilian coast during 1950–2021.



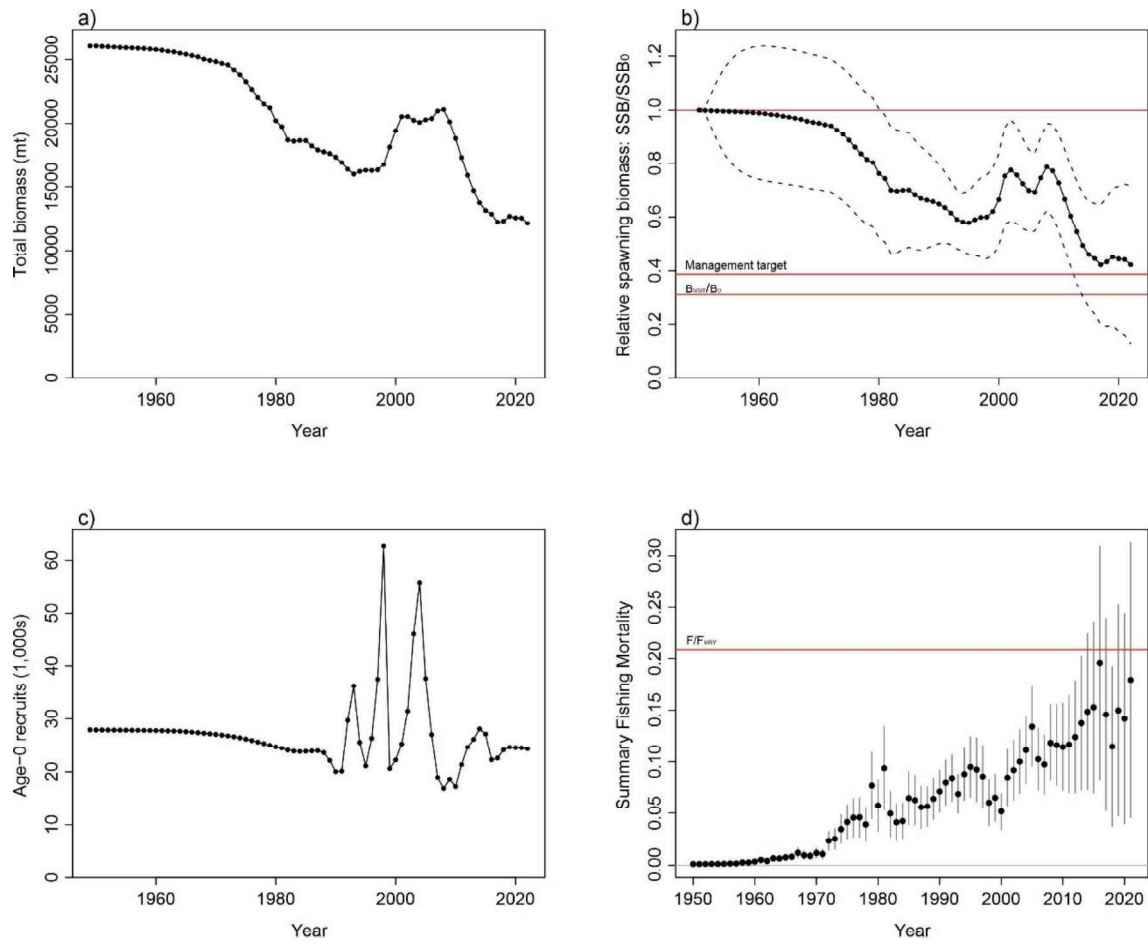
2243

2244 Figure 99: Spawning stock biomass (SSB) (a), relative spawning biomass SSB/SSB_0 (red
 2245 horizontal line represents relative SSB at MSY and the management goal of 25% above MSY)
 2246 (b), age-0 recruitment (c), and fishing mortality (red horizontal line represents fishing mortality
 2247 at MSY) (d) estimated by the Stock Assessment Continuum Tool for *Lutjanus jocu* along the
 2248 northeastern Brazilian coast during 1950–2021.



2249

2250 Figure 1010: Trajectory of stock spawning biomass (SSB) (a), relative spawning biomass
 2251 SSB/SSB₀ (red horizontal line representing the relative SSB at the MSY and the management
 2252 goal of 25% above MSY) (b), age-0 recruitment (c), and fishing mortality (red horizontal line
 2253 representing the fishing mortality at the MSY) (d) estimated by the Stock Assessment
 2254 Continuum Tool for *Lutjanus synagris* in the Northeast Brazil between 1950 and 2021.



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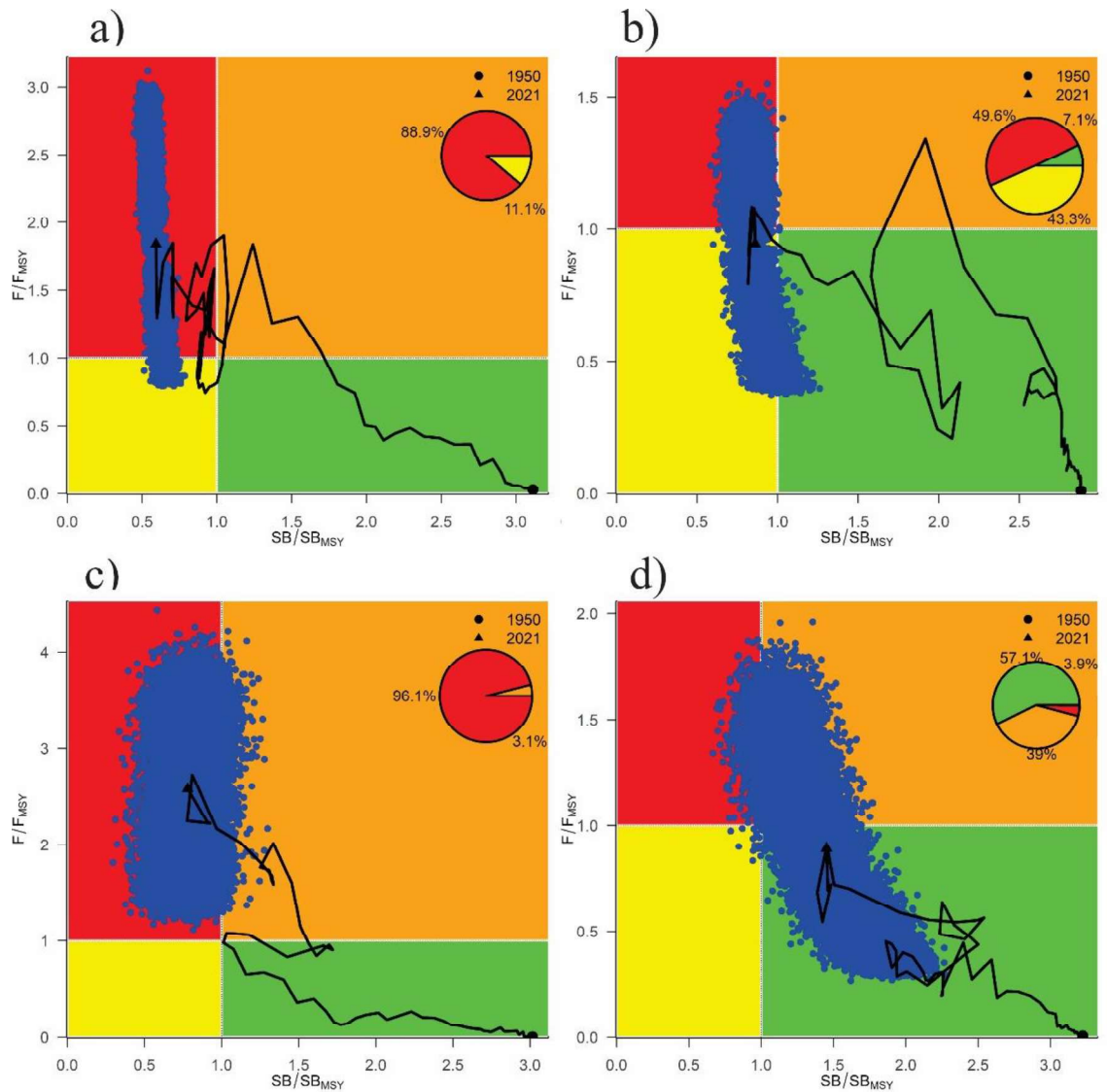
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Figure 1111: Trajectory of stock spawning biomass (SSB) (a), relative spawning biomass SSB/SSB₀ (red horizontal line representing the relative SSB at the MSY and the management goal of 25% above MSY) (b), age-0 recruitment (c), and fishing mortality (red horizontal line representing the fishing mortality at the MSY) (d) estimated by the Stock Assessment Continuum Tool for *Ocyurus chrysurus* in the Northeast Brazil between 1950 and 2021.



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2262

2263 Figure 12: Relative fishing mortality (F) in relation to relative spawning stock biomass (SSB) of

2264 a) *Lutjanus analis*, b) *Lutjanus jocu*, c) *Lutjanus synagris*, and d) *Ocyurus chrysurus* sampled

2265 along the northeastern Brazilian coast during 1950–2021. Red (top left) corresponds to

2266 “overfished and under overfishing”; green panel (bottom right) corresponds to “under no risk”;

2267 yellow (lower left) corresponds to “overfished”; and orange (upper right) corresponds to “under

2268 overfishing”. Blue dots represent individual bootstrap runs for each combination of randomly

2269 sampled F and SSB . Pie charts indicate the probability of each species falling within each risk

category.

2270

Tables

2271

2272 Table 1: Input parameters used in the Stock Assessment Continuum Tool reference model for a) *Lutjanus analis*, b) *Lutjanus jocu*, c) *Lutjanus synagris*, and d)

2273 *Ocyurus chrysurus* sampled along the northeastern Brazilian coast during 1950–2021 (sources: 1 – Lessa, Bezerra Junior and Nóbrega 2004; 2 – Previero et al. 2011; 3 – Schwaborn et al. 2023, 4 – Araújo, Martins and Costa 2002).

Parameter	<i>Lutjanus analis</i>			<i>Lutjanus jocu</i>			<i>Lutjanus synagris</i>			<i>Ocyurus chrysurus</i>		
	Value	CI	Source	Value	CI	Source	Value	CI	Source	Value	CI	Source
Maximum age A_{max}	29	-	1	29	-	1	22	-	1	19	-	1
Mean age at 50% maturity A_{50}	2.27	-	1	2.89	-	1	1.81	-	1	2.24	-	1
VBGF Asymptotic length L_{∞}	92.3	-	Fishlife	87.8	-	2	59.7	-	3	56.7	-	4
VBGF growth coefficient k_{yr}^{-1}	0.16	-	Fishlife	0.10	-	2	0.20	-	3	0.13	-	4
VBGF Age at length 0 t_0	NA	-	Fishlife	-1.49	-	2	NA	-	3	-0.77	-	4
Mean length at 50% maturity L_{50}	28.00	-	1	32.4	-	1	18.1	-	1	20.1	-	1
Mean length at 95% maturity L_{95}	34.00	-	1	36.4	-	1	23.3	-	1	26.0	-	1
Natural Mortality M	0.21	0.17 – 0.25	-	0.22	0.19 – 0.25	-	0.24	0.20 – 0.28	-	0.29	0.25 – 0.34	-
WL relationship - α	0.02	-	1	0.02	-	1	0.01	-	1	0.03	-	1
WL relationship - β	2.96	-	1	2.97	-	1	3.08	-	1	2.74	-	1
Weight-based fecundity coefficient	0.02	-	-	0.02	-	-	0.01	-	-	0.03	-	-
Weight-based fecundity exponent	2.96	-	-	2.97	-	-	3.08	-	-	2.74	-	-
Steepness h	0.70	-	-	0.70	-	-	0.70	-	-	0.70	-	-

2274

2275 Table 2: Spawning stock biomass (SSB), catch at maximum sustainable yield (MSY), stock biomass at MSY, SSB in 2021 in relation to SSB at MSY, fishing
 2276 mortality at MSY, fishing mortality in 2021 in relation to fishing mortality at MSY, and stock biomass status estimated with the Stock Assessment Continuum
 2277 Tool for *Lutjanus analis*, *Lutjanus jocu*, *Lutjanus synagris*, and *Ocyurus chrysurus* sampled along the northeastern Brazilian coast during 1950–2021.

Parameter	<i>Lutjanus analis</i>		<i>Lutjanus jocu</i>		<i>Lutjanus synagris</i>		<i>Ocyurus chrysurus</i>		Dimension
	Estimates	SD	Estimates	SD	Estimates	SD	Estimates	SD	
Spawning stock biomass unfished	7,500	236	3,348	209	8,090	631	11,793	1,412	Tons
Catch at maximum sustainable yield (MSY)	638	20	403	25	807	63	1,941	228	Tons
Stock Spawning Biomass at MSY (SSB _{MSY})	2,409	76	1,162	72	2,683	210	3,659	433	Tons
SSB ₂₀₂₁ /SSB _{MSY}	0.586	-	0.814	-	0.779	-	1.422	-	-
Fishing mortality at MSY (F _{MSY})	0.126	0.001	0.146	0.001	0.146	0.001	0.208	0.002	Year ⁻¹
F ₂₀₂₁ /F _{MSY}	1.831	-	1.002	-	2.629	-	0.859	-	-
Stock biomass status	Overexploited and under overexploitation		Overexploited and under slight overexploitation		Overexploited and under heavy overexploitation		Underexploited and not under overexploitation		-

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Considerações finais

A exploração pesqueira no Brasil em geral é historicamente marcada pela limitação robustos, dificultando tanto a avaliação dos estoques quanto a implementação de políticas adequadas de manejo. Nesse contexto, as questões centrais deste estudo foram: (a) a viabilidade de modelos para dados limitados como alternativa métodos tradicionais e (b) a necessidade de atualização do estado dos estoques das principais espécies de lutjanídeos exploradas comercialmente. A partir de uma abordagem metodológica atual, esta tese buscou responder a essas questões e contribuir para o fortalecimento da gestão sustentável da pesca no Brasil, sendo seus resultados apresentados em dois artigos científicos que compõem o núcleo do trabalho.

O primeiro artigo avaliou o desempenho de modelos para dados limitados em comparação com abordagens tradicionais considerando pescarias tropicais de pequena escala. Diferentes configurações foram testadas utilizando o Stock Assessment Continuum Tool, escolhido por permitir a execução de avaliações sob diferentes níveis de disponibilidade de dados dentro de uma mesma estrutura de modelagem. Os resultados indicaram que, embora dependam de pressupostos mais restritivos e apresentem maior incerteza, esses modelos são capazes de fornecer diagnósticos úteis sobre o estado dos estoques em contextos onde os dados são escassos ou fragmentados. Entre as abordagens disponíveis, modelos baseados em comprimento mostram-se particularmente adequados para pescarias artesanais, uma vez que se utilizam de dados relativamente simples e baratos de coletar e permitem inferências objetivas sobre o estado do estoque (Medeiros-Leal et al., 2023). Em contraste, modelos baseados apenas em captura podem ser úteis para estimativas de produtividade e limites de captura, mas dependem de pressupostos fortes sobre a relação entre captura e biomassa, frequentemente difíceis de sustentar em pescarias multiespecíficas e pouco monitoradas. Assim, a escolha do método deve sempre considerar a disponibilidade e qualidade das informações, bem como a biologia das espécies avaliadas.

O segundo artigo apresentou uma atualização do estado dos estoques das quatro principais espécies de lutjanídeos exploradas no Nordeste do Brasil, representando a primeira avaliação abrangente dessas espécies em aproximadamente duas décadas. Utilizando dados de captura, composições de comprimento, índices de abundância e parâmetros de produtividade dentro da estrutura do SAC Tool, foi possível estimar biomassa e mortalidade relativas mesmo em um contexto de dados limitados. Os

resultados indicam que três das quatro espécies avaliadas apresentam elevada probabilidade de se encontrarem simultaneamente sobrepescadas e sob sobrepesca, enquanto *Ocyurus chrysurus* apresenta situação relativamente mais confortável, embora ainda sujeita a risco caso o esforço de pesca siga aumentando. Esses resultados evidenciam a necessidade de medidas de manejo capazes de reduzir a pressão pesqueira e favorecer a recuperação dos estoques.

A discussão integrada dos dois estudos reforça o papel estratégico dos modelos para dados limitados em regiões como o Nordeste do Brasil, onde limitações logísticas e financeiras dificultam a coleta contínua de informações pesqueiras. Mesmo sob condições adversas, é possível gerar estimativas consistentes que subsidiem decisões de manejo, embora a qualidade das avaliações dependa diretamente da melhoria contínua dos sistemas de monitoramento.

Os resultados obtidos possuem implicações diretas para a gestão pesqueira nacional. A condição crítica observada para parte das espécies indica a urgência de ações voltadas à redução da mortalidade por pesca, incluindo medidas como restrições sazonais, áreas de proteção marinha e instrumentos de planejamento espacial marinho. Igualmente importante é a integração da comunidade pesqueira no processo de tomada de decisão, uma vez que o sucesso de qualquer política de manejo depende do envolvimento e da aceitação dos principais atores. A abordagem adotada nesta tese, baseada na utilização de uma estrutura integrada capaz de acomodar diferentes tipos e níveis de dados, oferece ainda um caminho replicável para avaliação de outras espécies com características ecológicas semelhantes, contribuindo para a paronização e ampliação de avaliações em regiões com limitação de informações.

Conclui-se, portanto, que esta tese não apenas respondeu às questões científicas inicialmente propostas, como também demonstrou a aplicabilidade prática de modelos para dados limitados na avaliação e gestão de pescarias tropicais. Ao combinar avanços metodológicos com diagnósticos atualizados do estado dos estoques, o estudo contribui para o desenvolvimento de estratégias mais sustentáveis de exploração dos recursos marinhos. Mais do que isso, evidencia a necessidade de um compromisso contínuo com a coleta de dados, o aprimoramento das políticas de manejo e a integração entre ciência, gestão e sociedade, elementos essenciais para garantir a sustentabilidade ecológica e socioeconômica das pescarias e das comunidades que elas dependem.

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